Integrated Monitoring Plan for the Hanford Groundwater Monitoring Project

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October 2000

Prepared for the U.S. Department of Energy under Contract DE-AC06-76RL01830

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Summary

Groundwater is monitored at the Hanford Site to fulfill a variety of state and federal regulations, including the *Atomic Energy Act of 1954*; the *Resource Conservation and Recovery Act of 1976*; the *Comprehensive Environmental Response, Compensation, and Liability Act of 1980*; and Washington Administrative Code. Separate monitoring plans are prepared for various requirements, but sampling is coordinated and data are shared among users to avoid duplication of effort. The U.S. Department of Energy manages these activities through the Hanford Groundwater Monitoring Project.

This document is an integrated monitoring plan for the groundwater project. It documents well and constituent lists for monitoring required by the *Atomic Energy Act of 1954* and its implementing orders; includes other, established monitoring plans by reference; and appends a master well/constituent/ frequency matrix for the entire site.

The objectives of monitoring fall into three general categories: plume and trend tracking, treatment/ storage/disposal unit monitoring, and remediation performance monitoring. Criteria for selecting *Atomic Energy Act of 1954* monitoring networks include locations of wells in relation to known plumes or contaminant sources, well depth and construction, historical data, proximity to the Columbia River, water supplies, or other areas of special interest, and well use for other programs. Constituent lists were chosen based on known plumes and waste histories, historical groundwater data, and, in some cases, statistical modeling. Sampling frequencies were based on regulatory requirements, variability of historical data, and proximity to key areas. For sitewide plumes, most wells are sampled every 3 years. Wells monitoring specific waste sites or in areas of high variability will be sampled more frequently.



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1.0 Introduction

Groundwater is monitored in hundreds of wells at the Hanford Site to fulfill a variety of requirements. Separate monitoring plans are prepared for various requirements, but sampling is coordinated and data are shared among users to avoid duplication of effort. The U.S. Department of Energy (DOE) manages these activities through the Hanford Groundwater Monitoring Project (groundwater project), which is the responsibility of Pacific Northwest National Laboratory. The groundwater project does not include all of the monitoring to assess performance of groundwater remediation or all monitoring associated with active facilities.

This document is an integrated monitoring plan for the groundwater project and contains well and constituent lists for monitoring required by the *Atomic Energy Act of 1954* and its implementing orders (surveillance monitoring); other, established monitoring plans by reference; and a master well/constituent/frequency matrix for the entire Hanford Site.

1.1 Purpose

The purpose of this plan is to provide a venue to integrate groundwater monitoring for various requirements on the Hanford Site. Specific objectives of this plan are the following:

- design and describe monitoring well networks, constituent lists, sampling frequency, and quality assurance/quality control for the surveillance monitoring network; explain criteria used to design the program
- encompass Resource Conservation and Recovery Act of 1976 (RCRA), Comprehensive Environmental Response, Compensation, and Liability Act of 1980 (CERCLA), Washington Administrative Code (WAC) regulations, and other monitoring plans by reference
- provide well, constituent, and sampling frequency lists for all groundwater monitoring on the site.

This plan is subordinate to the *Environmental Monitoring Plan*, *U.S. Department of Energy, Richland Operations Office* (DOE 1997a), which is required by DOE Orders, and the *Hanford Site Ground-Water Protection Management Plan* (Barnett et al. 1995). This plan describes how DOE will implement the groundwater monitoring requirements described in those documents.

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¹ Pacific Northwest National Laboratory is operated by Battelle for DOE.

1.2 Objectives of Groundwater Monitoring

The Environmental Monitoring Plan (DOE 1997a) lists the purposes and objectives of groundwater monitoring and the groundwater project. These purposes and objectives fall into three general categories: 1) plume and trend tracking, 2) monitoring of treatment/storage/disposal units, and 3) independent assessment of performance monitoring for groundwater remediation activities (Table 1.1).

Plume and trend tracking are the primary objectives of surveillance monitoring. Treatment/storage/disposal unit monitoring includes units regulated under RCRA or state codes (recently active sites), CERCLA (past-practice sites), and the *Atomic Energy Act of 1954*. Monitoring associated with remediation activities is the responsibility of the environmental restoration contractor, but the groundwater project is responsible for "providing continuing, independent assessment of groundwater remediation activities" (DOE 1997a).

Table 1.1. Objectives of Groundwater Monitoring (DOE 1997a)

Plume and Trend Tracking

Determine baseline conditions of groundwater quality and quantity.

Characterize and define hydrogeologic, physical, and chemical trends in the groundwater system.

Identify existing and potential groundwater contamination sources.

Assess existing and emerging groundwater quality problems.

Evaluate existing and potential offsite impacts of groundwater contaminants.

Provide data on which decisions can be made concerning land-disposal practices and management and protection of groundwater resources.

Treatment/Storage/Disposal Unit Monitoring

Demonstrate compliance with applicable regulations and orders (RCRA, WAC).

Provide data to permit early detection of groundwater pollution or contamination.

Groundwater Remediation Performance Monitoring

Provide continuing, independent assessment of groundwater remediation activities (groundwater remediation and performance monitoring are conducted by the environmental restoration contractor; currently, Bechtel Hanford Inc.; groundwater project provides independent assessment).

1.3 Organization of This Plan

A brief overview of the hydrogeology of the Hanford Site is provided in Chapter 2 as background for the remainder of the plan. Chapter 3 describes the monitoring program, with an explanation of criteria for choosing well networks, constituent lists, and sampling frequency. Chapters 4 through 9 describe the waste sites, monitoring history, and a conceptual model of the movement of contaminants for each geographic region of the site. Chapter 10 describes the sampling and analysis plan, including methods for sampling and analysis, quality assurance, and quality control. Chapter 11 describes the water-level monitoring program; Chapter 12 describes data management, compliance issues, and reporting; followed by Chapter 13, the references cited herein.

An integrated monitoring matrix is presented in Appendix A, showing the wells to be sampled in fiscal year 2001. The appendix is updated annually.

2.0 Hydrogeology

The hydrogeology of the Hanford Site has been described in documents such as Hartman 2000, Chapter 3. A brief summary is provided here for the reader's convenience.

The uppermost aquifer beneath most of the Hanford Site is unconfined and composed of unconsolidated to semiconsolidated sands and gravels deposited on basalt bedrock. In some areas, deeper parts of the aquifer are locally confined by layers of silt and clay. Confined aquifers occur within the underlying basalt flows and associated sedimentary interbeds. A simplified stratigraphic column is illustrated in Figure 2.1.

Groundwater in the unconfined aquifer system generally moves from recharge areas along the western boundary of the site to the east and north toward the Columbia River, which is the major discharge area. This natural flow pattern was altered by the formation of groundwater mounds created by large volumes of artificial recharge at wastewater-disposal facilities. These mounds are declining, and groundwater flow is gradually returning to earlier patterns. Figure 2.2 shows a water-table map for March 1999.

The extent of major radionuclide contaminants in groundwater in 1999 is illustrated in Figure 2.3. Iodine-129, strontium-90, technetium-99, tritium, and uranium were present at levels above drinking water standards. Carbon-14, cesium-137, and plutonium exceeded standards in smaller areas. The extent of major hazardous chemical constituents in 1999 is shown in Figure 2.4. The most significant of these include carbon tetrachloride, chromium, and nitrate. Arsenic, fluoride, and trichloroethylene are also elevated in smaller areas.

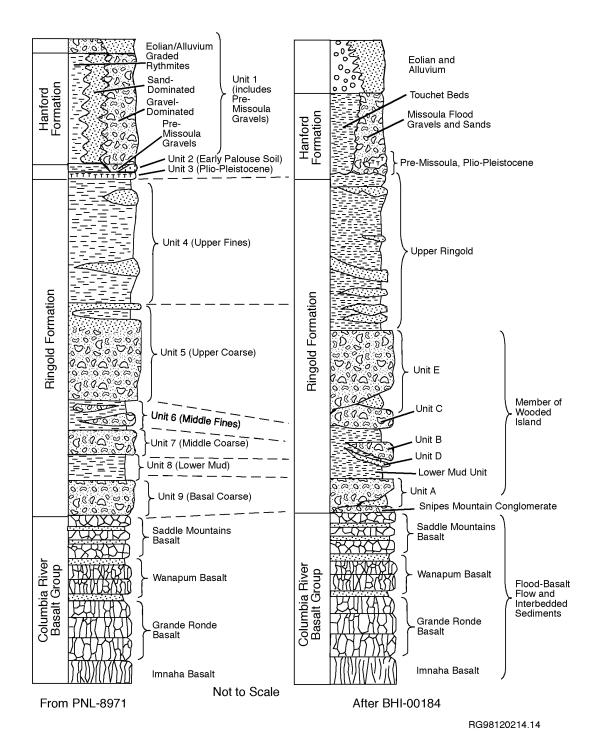


Figure 2.1. Comparison of Generalized Hydrogeologic and Geologic Stratigraphy

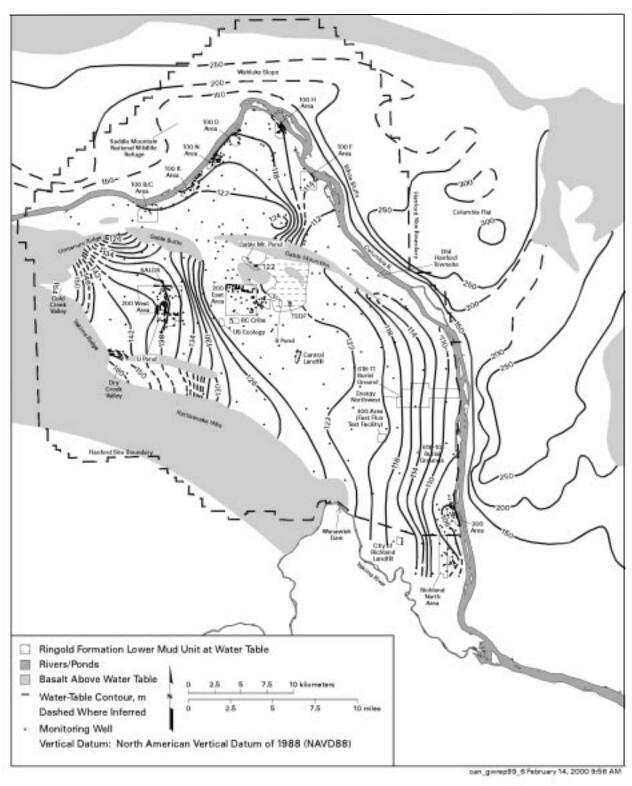


Figure 2.2. Hanford Site and Outlying Areas Water-Table Map, March 1999 (from Hartman et al. 2000)

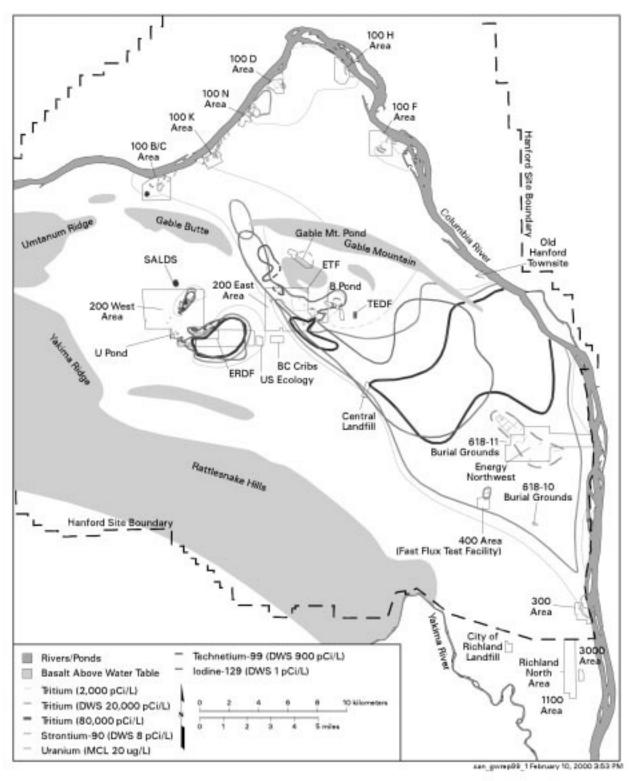


Figure 2.3. Distribution of Major Radionuclides in Groundwater at Concentrations Above Maximum Contaminant Levels or Interim Drinking Water Standards, Fiscal Year 1999 (Hartman et al. 2000)

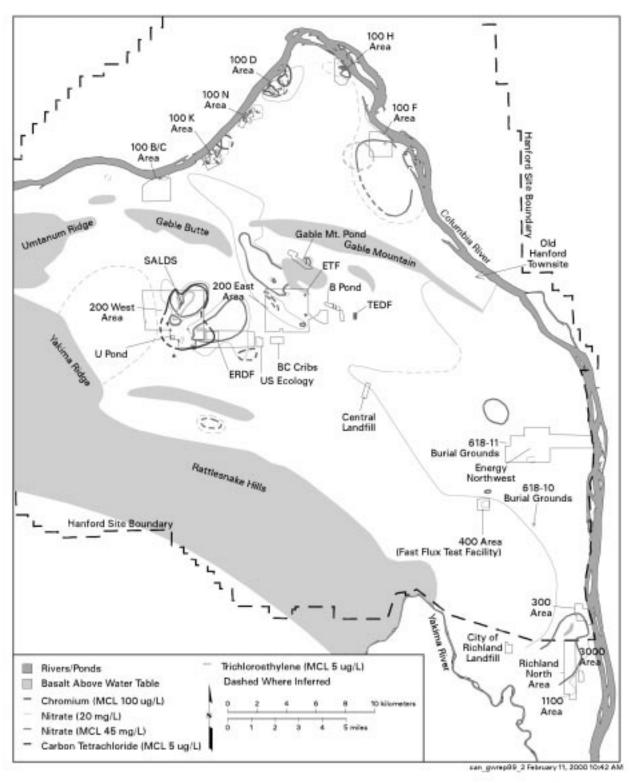


Figure 2.4. Distribution of Major Hazardous Chemicals in Groundwater at Concentrations Above Maximum Contaminant Levels, Fiscal Year 1999 (Hartman et al. 2000)

3.0 Monitoring Program

The integrated sampling and analysis matrix for the groundwater project is given in Appendix A. The matrix was designed for use in fiscal year 2001, but also includes wells that will be sampled every 2 or 3 years, as discussed in Section 3.3. The matrix includes well name, program, project, sampling frequency, and constituents to be monitored. Additional details, such as schedule and analytical methods, reside in a project database.

3.1 Groundwater Monitoring Network

Wells on the Hanford Site are monitored in compliance with: 1) the *Atomic Energy Act of 1954* and its implementing orders ("surveillance monitoring"), 2) CERCLA operable units, 3) remedial action performance assessment, 4) RCRA, and 5) WAC permits. Monitoring networks for items 2 through 5 are defined in monitoring plans, interim records of decision, or change agreements listed in Chapters 4 through 8. These monitoring networks are included in the monitoring matrix of Appendix A. The criteria for choosing wells for surveillance monitoring are discussed below.

- Defining plumes A representative areal distribution of wells within the plume is monitored, with an
 emphasis on wells with the highest concentrations of contaminants and wells near plume boundaries.
 Some wells in uncontaminated areas between plumes are also monitored to help control interpretation
 of plume boundaries and to monitor plume migration. Plumes migrating onto the site from offsite
 sources are also monitored (e.g., agricultural effects, Richland Landfill, Siemens Power Corporation).
 A geostatistical approach was employed to determine which wells should be sampled to track major
 plumes from the 200 West and 200 East Areas (discussed in Chapters 5 and 6).
- 2. Monitoring contaminant sources Waste-disposal facilities not regulated by RCRA or the WAC are included in surveillance monitoring (e.g., 100-K basins, 216-U-1 crib). Wells downgradient of these facilities are monitored to detect their impact on groundwater.
- 3. Interval monitored Most of the groundwater contamination on the Hanford Site is contained in the uppermost (unconfined) aquifer, so most of the monitoring wells are screened there. Newer wells installed for RCRA and CERCLA are screened across the water table and monitor the top 3 to 10 meters of the unconfined aquifer. Wells that monitor a longer interval are less desirable because contaminants could be diluted from representative concentrations to below detection limits. A few wells monitor deeper intervals of the suprabasalt sediments or confined aquifers in the basalt. These wells are sampled to monitor whether contamination has migrated deeper in the hydrologic system.

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¹ This type of monitoring has been called "operational monitoring" in the past. It is now considered part of surveillance monitoring.

- 4. Historical data Previous groundwater chemistry or water-level data in a well are useful for monitoring trends and for determining sampling frequency and constituent lists. Wells with historical data are preferable to those without.
- 5. Adequacy of well construction Wells with poor seals, broken casing, or other problems may not provide representative data, and will be remediated or decommissioned.
- 6. Amount of water in the well Declining water levels are causing some wells to go dry. Wells that are likely to contain sufficient water for sampling are chosen for the network.
- 7. Proximity to the Columbia River In some cases, it is desirable to monitor wells very near the river shore to assess what concentrations of contaminants are entering the river. In other cases, it is more advantageous to choose wells farther inland to avoid fluctuations in concentration caused by bank storage effects.
- 8. Use by other requirements of the groundwater project (e.g., RCRA, CERCLA) Wells being sampled for other purposes are used for surveillance monitoring, where possible, for a more cost-effective program.
- 9. "Guard wells" Key areas have been identified as being of special interest: bands of wells in Gable Gap and southeast of the 200 East Area were chosen to monitor contamination migrating out of the 200 Areas (discussed in Chapter 6), wells near the Columbia River, wells in the southern portion of the site near the city of Richland's North Well Field and recharge basins.
- 10. Performance assessment The environmental restoration contractor (i.e., Bechtel Hanford, Inc.) conducts performance assessment monitoring in conjunction with remedial actions. The groundwater project is responsible for providing independent assessment of remedial actions, so wells near the remedial actions are included.

3.2 Constituents

Constituents are included in the sampling matrix of Appendix A. This matrix is an abbreviated version of the sampling matrix maintained by the groundwater project, which specifies various methods of analysis for some constituents.

The following criteria were considered to determine what analyses should be run on the samples for surveillance monitoring:

- 1. Proximity to known plumes or waste sites If a well is located in a contaminant plume or downgradient of a plume, it is generally sampled for that contaminant.
- 2. Historical data in well Wells are generally not sampled for constituents that have not been detected or are below some level of interest (e.g., drinking water standards) unless they are monitoring movement of a nearby plume.

- 3. Statistical modeling (discussed in Chapter 6).
- 4. Use for other requirements If there is a choice of analytical method for a desired constituent, the method used for other monitoring purposes is chosen if it is satisfactory for surveillance monitoring.
- 5. Washington State Department of Health constituents Constituents, including total alpha, anions, total beta, gamma, iodine-129, technetium-99, tritium, and uranium isotopes, are co-sampled to provide a quality control check.

The choice of constituents for RCRA, CERCLA, and other monitoring requirements are based on waste history, permit conditions, and constituents of concern, as discussed in their monitoring plans.

3.3 Sampling Frequency

Sampling frequency for RCRA, CERCLA, and other monitoring requirements are determined by regulation, permits, or other agreements. Frequency for plume and trend tracking are based on the following criteria:

- 1. Variability of historical data If previous concentrations are level or are on a steady trend, less-frequent sampling (every 3 years) is sufficient. Wells with larger variability are sampled more frequently (annually or more often).
- 2. Proximity to key areas Guard wells (see Section 3.1) and wells monitoring source areas are sampled more frequently.
- 3. Mobility of contaminants in groundwater Contaminants with greater mobility (e.g., tritium) may be sampled more frequently than those that are not very mobile in groundwater (e.g., strontium-90).

3.4 Changes to Monitoring Program

As data are received and evaluated, changes will be made to the program, as needed. For example, if the concentration of a contaminant in a well increases suddenly, an additional sample may be collected and analyzed to confirm or refute the initial result. This type of "one-time" change may be made without revision of this plan.

Each year the well/constituent matrix in this plan will be reviewed for adequacy and revised, if necessary, for the following fiscal year. These revisions will incorporate any changes made to monitoring plans for RCRA, CERCLA, and other requirements.

4.0 100 Areas

For the purposes of this plan, "100 Areas" describes that portion of the Hanford Site north of Gable Mountain and Gable Butte and south of the Columbia River and includes the six reactor areas (B/C, K, N, D, H, and F [upstream to downstream]) and the 600 Area in between.

4.1 Background

4.1.1 Waste Sites, Discharges, and Groundwater Operable Units

Hundreds of waste sites have been identified in the 100 Areas, including fuel storage or retention basins that leaked; effluent disposal cribs, ditches, and drains; and various spills or other unplanned releases. Those with site-specific monitoring requirements and those that appear to have affected groundwater quality are listed in Table 4.1.

Most of the sites are inactive radiological or mixed waste sites, and are being cleaned up or monitored under the requirements of CERCLA or as RCRA past-practice sites. Four sites are regulated under RCRA because they were more recently active and contained dangerous waste constituents. Another RCRA site, 120-D-1 ponds, was "clean-closed," and no longer requires monitoring. Two sites currently discharge nondangerous effluent to the ground (sanitary waste and filter backwash in the 100 N Area).

Groundwater beneath the reactor areas and surrounding areas is divided into five groundwater operable units: 100-BC-5 (100 B/C Area), 100-KR-4 (100 K Area), 100-NR-2 (100 N Area), 100-HR-3 (100 D and 100 H Areas), and 100-FR-3 (100 F Area). Pump-and-treat systems are active in the 100 K, 100 D, and 100 H Areas for chromium and in the 100 N Area for strontium-90. An in situ treatment system is active in the 100 D Area to chemically reduce hexavalent chromium to insoluble chromium compounds (redox manipulation). All of these remediation systems are considered interim actions; final remedial actions have not yet been selected.

4.1.2 Groundwater Monitoring Requirements and History

Limited groundwater monitoring has been conducted in the reactor areas since the 1940s. Very few monitoring wells existed in the early decades but more were installed in the 100 K, 100 N, and 100 H Areas in the 1970s and monitored for DOE requirements. RCRA monitoring began in the late 1980s in the 100 N and 100 H Areas, and in the early 1990s in the 100 D Area, so additional wells were installed. CERCLA investigations and cleanup actions in the 1990s resulted in the installation of dozens more wells, spread among the reactor areas and the intervening 600 Area between the 100 D and 100 H Areas.

CERCLA interim actions in the 100 K, 100 N, 100 D, and 100 H Areas include specific monitoring requirements. CERCLA operable unit monitoring networks have also been defined for these areas and for

Table 4.1. Selected Waste Sites in the 100 Areas^(a)

		C ii i CI i C		
Facility (period of use)	Waste Type	Constituents of Interest for Groundwater Monitoring	Status	
	100 B/C Are	a		
116-B-11 (1944-68) and 116-C-5 (1952-69) retention basins	Reactor coolant effluent; leaks known	Radionuclides, metals strontium-90, chromium	Past-practice; contaminated soil removed	
116-B-1 (1950-68) and 116-C-1 (1952-68) waste disposal trenches	Coolant effluent from fuel- element failure (highly radioactive)	Radionuclides	Past-practice; contaminated soil removed	
116-B-5 crib (1950-68)	Process effluent	Tritium	Past-practice; contaminated soil removed	
118-B-6 burial ground (1950-53)	Contaminated equipment	High-level tritium	Past-practice	
Storage tanks and transfer facilities (1944-69)	Sodium dichromate leakage from water-treatment facilities	Chromium	Past-practice	
	100 K Area			
Reactor buildings fuel storage basins (KE: 1955-71; 1975-present. KW: 1955-71; 1981- present)	Radionuclide-contaminated water; leaks known	Tritium, strontium-90	Active	
116-KE-3 (1955-71) and 116-KW-2 (1955-70) french drain/reverse well	Effluent from fuel-storage basin drainage collection	Tritium, strontium-90	Past-practice	
116-KE-1 (1955-71) and 116-KW-1 (1955-71) cribs	Reactor condensate	Tritium, carbon-14	Past-practice	
116-K-2 trench (1955-71)	Reactor coolant water, decontamination liquids	Chromium, strontium-90	Past-practice	
116-KW-3 (1954-70) and 116-KE-4 (1955-71) retention basins	Reactor coolant; leaks known	Radionuclides	Past-practice	
Storage tanks and transfer facilities	Sodium dichromate leakage from 183-KE and 183-KW water-treatment facilities	Chromium	Past-practice	
100 N Area				
1301-N liquid waste disposal facility (1963-85)	Reactor coolant	Strontium-90, tritium, minor hazardous constituents ^(b)	RCRA past-practice	
1325-N liquid waste disposal facility (1983-89)	Reactor coolant	Strontium-90, tritium, minor hazardous constituents ^(b)	RCRA past-practice	
1324-NA percolation pond (1986-90)	Treated demineralizer effluent	Sulfate, sodium, pH ^(b)	RCRA past-practice	
Fuel station	Fuel tank leaks confirmed	Hydrocarbons	Past-practice	

Table 4.1. (contd)

		Constituents of Interest for	
Facility (period of use)	Waste Type	Groundwater Monitoring	Status
N Reactor basins	Fuel-storage basins	Radionuclides	Inactive
183-N backwash discharge pond (1983-present)	Filter backwash	None	Active
124-N-10 sewage lagoon (1987-present)	Sanitary waste	Nitrate, coliform	Active; WAC permitted
	100 D Area		
116-D-7 (1944-67) and 116-DR-9 (1950-67) retention basins	Reactor coolant; leaks known	Radionuclides, chromium	Past-practice; contaminated soil removed
116-D-1 (1947-67) and 116- DR-2 (1950-67) trenches	Highly radioactive coolant from fuel element failure	Radionuclides	Past-practice; contaminated soil removed
Reactor cribs, drains	Water and sludges from fuel storage basins; decontamination solutions; condensate from inert gas system	Carbon-14, nitrate, strontium-90	Past-practice
Storage tanks and transfer facilities	Sodium dichromate leakage from corrosion inhibitor	Chromium	Past-practice
120-D-1 ponds (1977-94)	Effluent from water treatment	pH, mercury ^(b)	Contaminated soil removed; RCRA clean closed ^(c)
	100 H Area		
116-H-7 (107-H) retention basin (1949-65)	Reactor coolant; leaks known	Tritium, strontium-90	Past-practice; contaminated soil removed
116-H-1 (107-H) trench (1952-65)	Highly radioactive coolant from reactor fuel-element failure	Tritium, strontium-90, nitrate	Past-practice; contaminated soil removed
Reactor cribs, drains	Water and sludge from fuel- storage basins; decontamination solutions	Chromium	Past-practice
183-H solar evaporation basins (1973-85)	Neutralized acid etch solutions	Technetium-99, uranium, nitrate, chromium, fluoride	RCRA, (d) basins and contaminated soil removed
	100 F Area		
116-F-14 retention basin and pipelines (1945-65)	Reactor coolant; leaks known	Strontium-90, chromium	Past-practice
116-F-2 trench (1950-65)	Highly radioactive coolant	Strontium-90, chromium	Past-practice
116-F-9 trench (1963-76)	Cleaning wastes from experimental animal laboratories	Radionuclides	Past-practice

Table 4.1. (contd)

Facility (period of use)	Waste Type	Constituents of Interest for Groundwater Monitoring	Status
116-F-3 (1947-51) and 116-F-6 (1952-65) trenches	Reactor coolant and sludge	Radionuclides	Past-practice
116-F-1 trench (1953-65)	Liquid waste from reactor and associated buildings	Radionuclides, metals, uranium, strontium-90, nitrate	Past-practice
118-F-1 (1954-65) and 118-F-6 (1965-73) solid waste-burial grounds	Contaminated equipment, animal wastes, coal ash	Tritium, plutonium	Past-practice

- (a) Sites with specific groundwater monitoring requirements and those that appear to have affected groundwater quality.
- (b) Known or suspected in waste; not significantly detected in groundwater to date.
- (c) Clean closed in 1999 (no waste left in place). No further RCRA monitoring required.
- (d) Groundwater beneath 183-H to be remediated under CERCLA.

the 100 B/C and 100 F Areas. The K basins, where spent reactor fuel rods are stored, have leaked in the past and are monitored under DOE Order 5400.1. Monitoring plans for the K basins, CERCLA, and RCRA are given in Table 4.2.

The *Atomic Energy Act of 1954* and DOE Order 5400.1 also require sitewide surveillance monitoring to track contaminant plumes. This document serves as the monitoring plan for surveillance monitoring performed per DOE orders.

4.2 Conceptual Model

The most widespread contaminants of concern in groundwater at the 100 Areas are hexavalent chromium, nitrate, and tritium. Groundwater is locally contaminated with carbone-14, strontium-90, sulfate, technetium-99, trichloroethylene, and uranium. Groundwater also flows into the 100 Areas through the gap between Gable Mountain and Gable Butte, carrying contamination from the 200 Areas.

Effluent from leaking retention basins and disposal trenches has contaminated the soil in the 100 Areas for ~50 years. Radionuclides with short half-lives decayed in the retention basins or in the vadose zone. Nonradioactive constituents and longer-lived radionuclides were carried down through the vadose zone beneath the waste sites. Some of these sorbed to sediment, some remained in the moisture in the vadose zone, and large quantities were carried into the groundwater (Figure 4.1).

When the reactors were active, huge volumes of water were discharged, creating large groundwater mounds that disrupted the natural groundwater flow. The contaminants moved outward on these mounds, contaminating a larger area in the saturated zone than in the vadose zone. The mounds dissipated after discharges ceased, and groundwater flow resumed its normal pattern (i.e., toward the river). Groundwater beneath the 100 Areas continues to carry contaminants to the river, where it discharges from springs,

Table 4.2. Groundwater Monitoring in the 100 Areas

Manitasina Danvinanant				
Monitoring Requirement (monitoring plan reference)	Comments			
100 B/C Area				
CERCLA (Federal Facility Agreement and Consent				
Order Change Control Form M-15-99-03; Sweeney	Long-term plume momtoring			
2000a)				
10	00 K Area			
CERCLA (ROD 1996a, DOE 1997b)	CERCLA interim action for chromium; wells near 116-K-2 trench ^(a)			
CERCLA (National Priorities List Change Control Form 108, 11/20/96)	100-KR-4 Operable Unit remedial investigation			
DOE Order 5400.1 (Johnson et al. 1995)	KE and KW fuel storage basins			
10	00 N Area			
RCRA (Hartman 1996)	1301-N, 1324-N/NA, 1325-N sites			
CERCLA (National Priorities List Change Control Form 113, 3/25/97)	N springs expedited response action (strontium-90 plume near 1301-N) ^(a)			
CERCLA (Federal Facility Agreement and Consent Order Change Control Form M-15-96-08, 10/9/96; Borghese et al. 1996)	100-NR-2 Operable Unit remedial investigation; also includes RCRA wells of Hartman (1996)			
10	00 D Area			
CERCLA (ROD 1996a, DOE 1997b)	CERCLA interim action for chromium; wells near retention basins and disposal trenches ^(a)			
CERCLA (ROD 1999)	100-HR-3 ROD amended to include in situ redox manipulation in southwestern 100 D Area.			
CERCLA (National Priorities List Change Control Form 107, 11/20/96)	100-HR-3 (D Area) Operable Unit remedial investigation			
10	00 H Area			
RCRA (Hartman 1997)	183-H solar evaporation basins			
CERCLA (ROD 1996a, DOE 1997b)	CERCLA interim action for chromium ^(a)			
CERLCA (Peterson and Raidl 1996; National Priorities List Change Control Form 107, 11/20/96)	100-HR-3 (H Area) Operable Unit remedial investigation			
100 F Area				
CERCLA (Federal Facility Agreement and Consent Order Change Control Form M-15-99-02; Sweeney 2000b)	Long-term plume monitoring			
(a) Groundwater monitored independently of groun(b) Will be eliminated when new permit issued.	dwater project.			

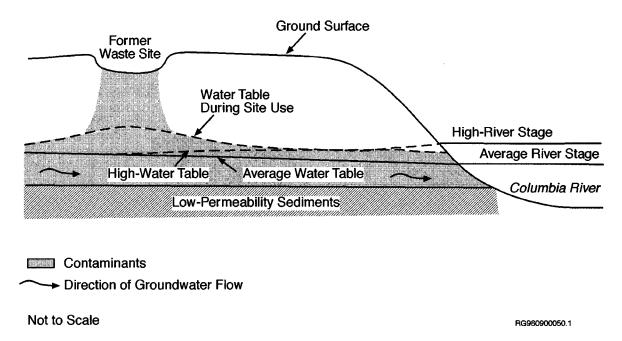


Figure 4.1. Conceptual Model of Subsurface Contamination in the 100 Areas

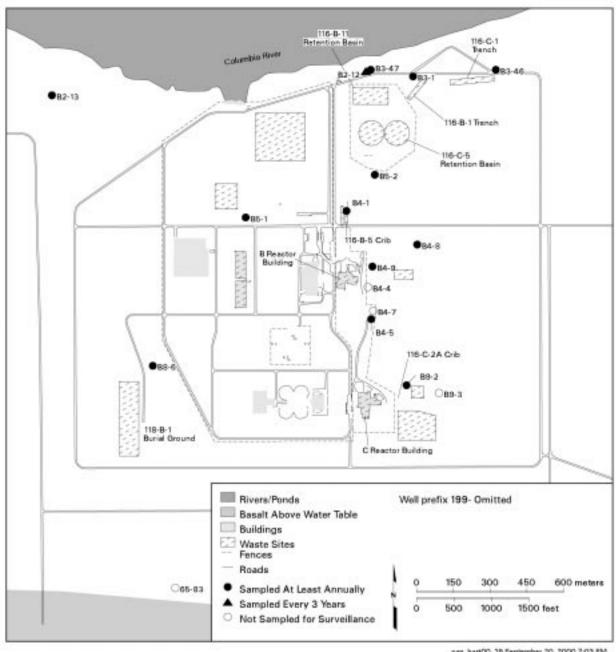
seeps, and through the riverbed below the water line. Groundwater nearest the river often has lower concentrations of contaminants because of dilution. When river stage is high, the water table may rise into the former mound areas and mobilize some constituents (see Figure 4.1) or it may dilute contaminants further. This influx of river water also temporarily disrupts the direction and rate of groundwater flow. Locally, groundwater extraction and injection also affect flow directions and intercept contaminants before they reach the river.

The vertical component of groundwater flow in the 100 Areas is generally upward, and most of the contamination is limited to the unconfined aquifer. However, it is likely that when groundwater mounds were present, there was a significant downward gradient, and several wells that monitor the confined Ringold or basalt-confined aquifers appear to be contaminated.

Contaminant concentrations are expected to decrease with time because of dispersion, dilution, radioactive decay, remediation, and discharge to the river. There are no new sources of contamination, but concentrations of existing contaminants will vary because of plume movement and mobilization of vadose zone contamination.

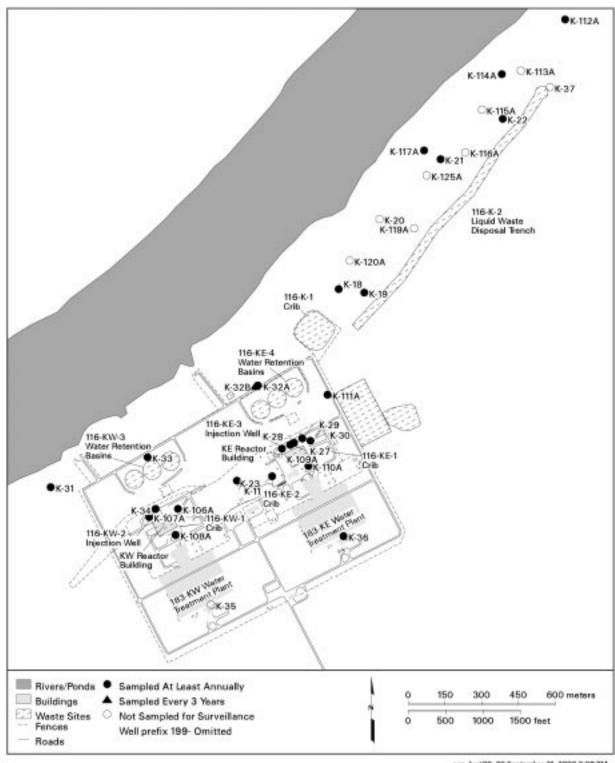
4.3 Monitoring Program

Locations of monitoring wells for the 100 Areas are illustrated in Figures 4.2 through 4.8. In addition to the shallow unconfined wells, the network includes most of the few available deeper wells (completed in the confined Ringold or the basalt-confined aquifer). Most of the 600 Area wells will be sampled every 3 years. Wells in the reactor areas are sampled every year, except for those wells near the river or wells with highly variable concentrations that are sampled more frequently. Wells monitoring chromium near the in situ redox manipulation application in the 100 D Area are monitored quarterly.



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Figure 4.2. Groundwater Project Well Locations: 100 B/C Area



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Figure 4.3. Groundwater Project Well Locations: 100 K Area

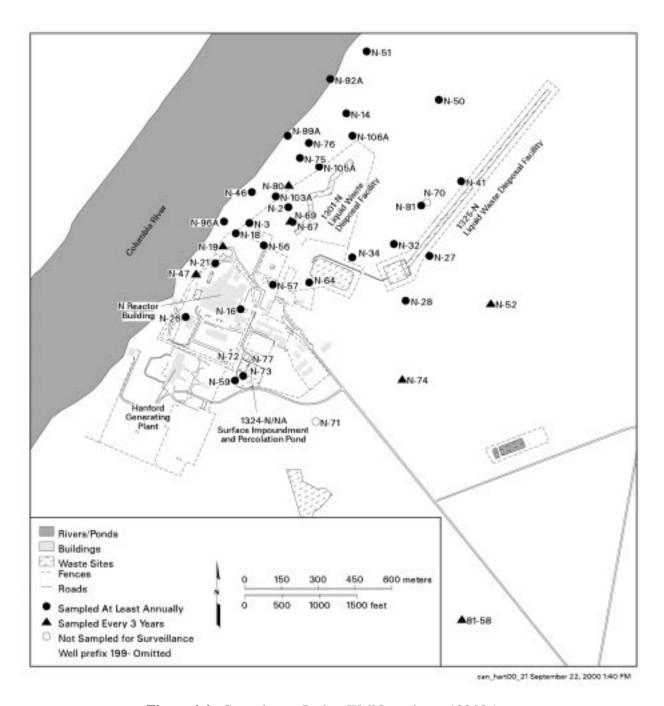


Figure 4.4. Groundwater Project Well Locations: 100 N Area

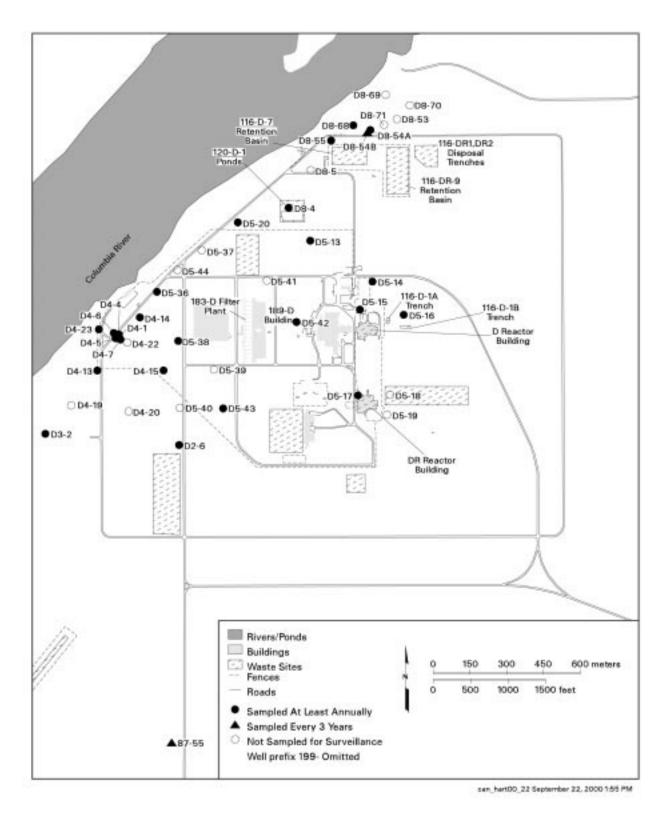


Figure 4.5. Groundwater Project Well Locations: 100 D Area

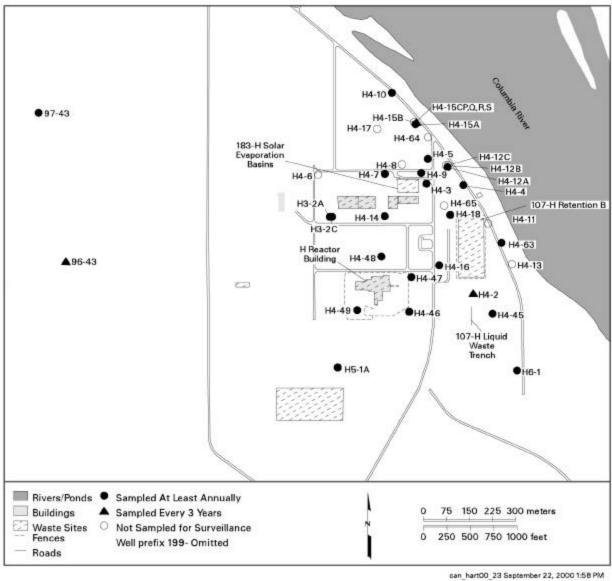
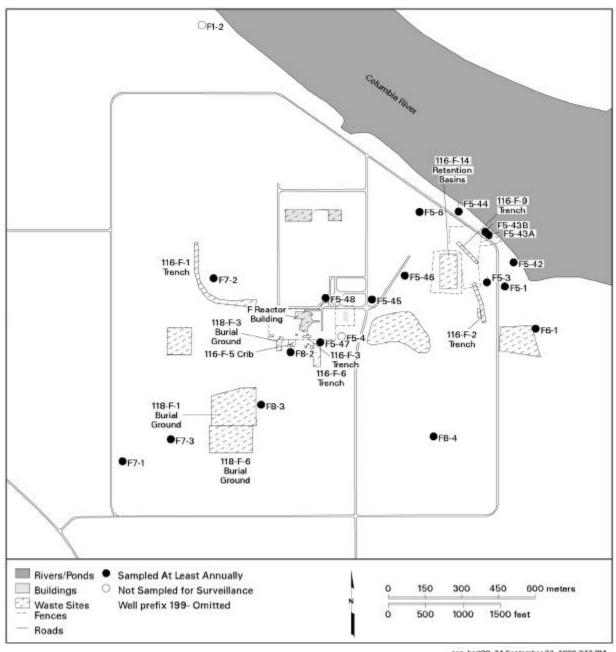


Figure 4.6. Groundwater Project Well Locations: 100 H Area



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Figure 4.7. Groundwater Project Well Locations: 100 F Area

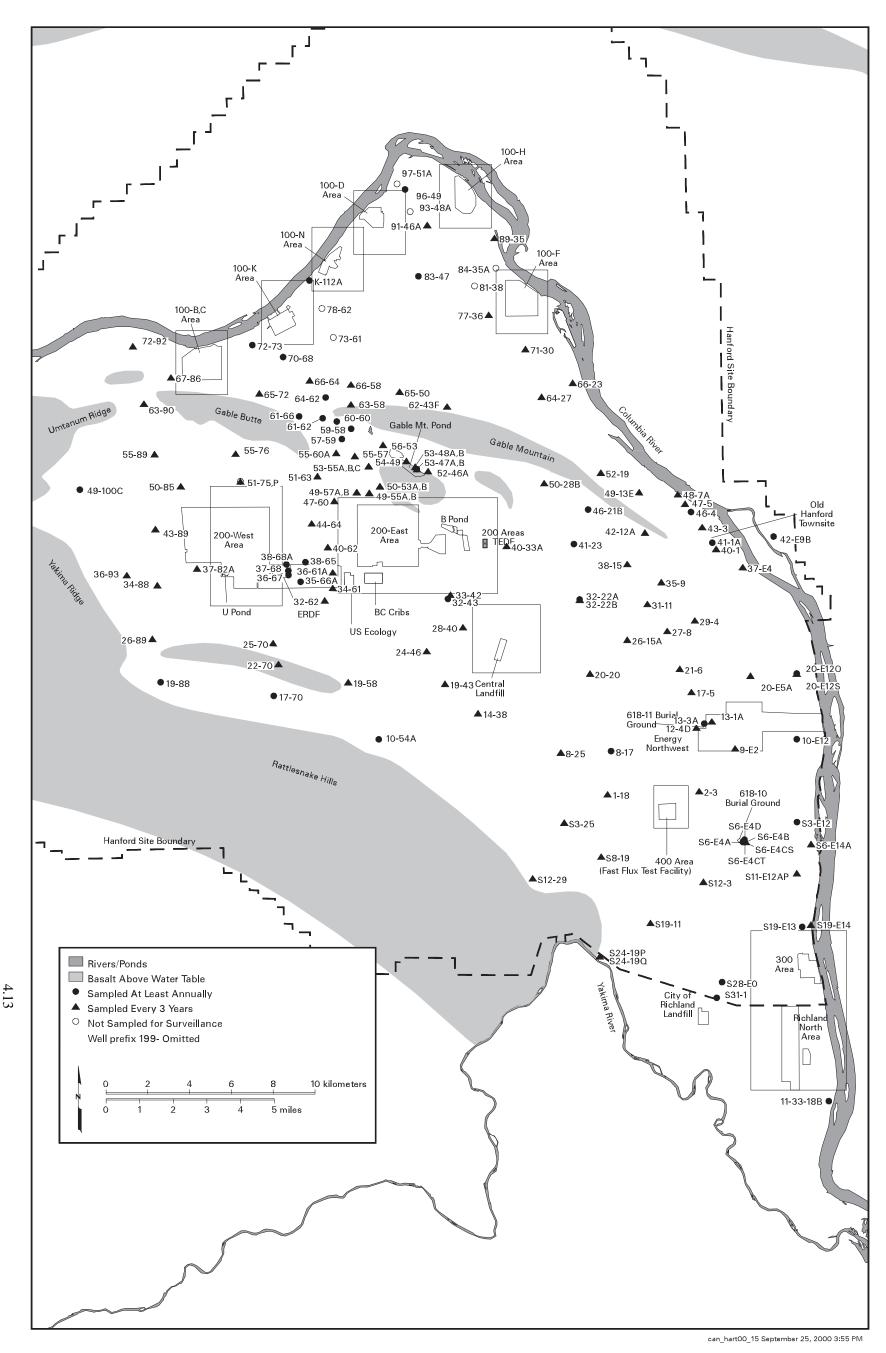


Figure 4.8. Groundwater Project Well Locations: 600 Area

5.0 200 West Area

The 200 West Area is located on the central plateau of the Hanford Site. Portions of the 600 Area affected by groundwater contamination originating in or near the 200 West Area were evaluated for the monitoring network in this chapter.

5.1 Background

Activities within this area have included irradiated nuclear fuel processing and liquid and solid waste storage and disposal. This area has been used since the 1940s.

5.1.1 Waste Sites, Discharges, and Groundwater Operable Units

Several processing facilities in the 200 West Area have contributed to groundwater contamination through disposal of radioactive and hazardous liquid wastes in ponds, cribs, ditches, and underground storage tanks. Large quantities of solid wastes, both from on and off the site, have been disposed of in numerous burial grounds in the 200 West Area. The sites with specific monitoring requirements and those that appear to have affected groundwater quality are listed in Table 5.1. Additional information is provided in Hartman (2000), and more complete site inventories are included in reports listed in the bibliography of that document. A number of facilities are regulated under RCRA because they were more recently active and contain, or contained, dangerous chemical waste constituents. Six RCRA units have groundwater monitoring requirements, including two low-level burial grounds, which are the only sites actively receiving waste within the 200 West Area. Four single-shell tank waste management areas located in the 200 West Area also are monitored under RCRA; the tanks currently are used to store mixed waste. The 616-A crib, also known as the State-Approved Land Disposal Site, is located just north of the 200 West Area. The site consists of a drain field that is used to dispose of liquid waste containing tritium.

Two CERCLA groundwater operable units relate to 200 West Area contamination. The 200-UP-1 Operable Unit generally covers the groundwater in the southeastern part of the area, where technetium-99 and uranium contamination near U Plant are being remediated by a CERCLA interim action. The 200-ZP-1 Operable Unit generally covers groundwater contamination originating in the northwestern part of the 200 West Area, where interim actions are in place to remediate carbon tetrachloride contamination.

5.1.2 Groundwater Monitoring Requirements and History

Groundwater monitoring wells in the 200 West Area were installed to monitor specific disposal facilities in the mid 1940s. RCRA monitoring wells were installed beginning in 1987. Several injection and extraction wells have been drilled to support interim action pump-and-treat systems. The *Atomic Energy Act of 1954* and DOE Order 5400.1 require monitoring to identify and track contaminant plumes. This document serves as the monitoring plan for surveillance monitoring per DOE orders.

Table 5.1. Selected Waste Sites in the 200 West Area^(a)

Facility	Waste Type	Constituents of Interest for Groundwater Monitoring	Type of Site (monitoring plan reference)
Single-shell tank farms (Waste Management Areas S-SX, T, TX-TY, and U)	Radioactive/chemical slurries	Sodium hydroxide, sodium salts, radionu- clides, ferrocyanide	RCRA (Johnson and Chou 1999; Caggiano and Chou 1993; ^(b) Hodges and Chou 2000)
T Plant disposal facilities (e.g., 216-T-26, -28, -19, -25)	Diverse chemical and radiological waste	Tritium, iodine-129, technetium-99, nitrate, chromium, carbon tetra- chloride, chloroform, trichloroethylene, fluoride	Past-practice (this document)
Reduction-Oxidation Plant disposal facilities (including 216-S-10 pond/ditch)	Solvent-extraction process wastes	Nitrate, trichloroethylene, tritium, iodine-129, technetium-99, uranium, strontium-90	Past-practice, except S-10: RCRA (Airhart et al. 1990)
U Plant disposal facilities (216-U-12 and other retention trenches)	Supernatant from scavenged waste	Iodine-129, technetium-99, uranium, nitrate, trichloroethylene	Past-practice, except U-12: RCRA (Jensen et al. 1990, Williams and Chou 1993). 200-UP-1 interim action (DOE 1997c, ROD 1997) ^(c)
Plutonium Finishing Plant disposal facilities (e.g., 216-Z-1A and -Z-9)	Transuranic and chemical wastes	Nitrate, carbon tetra- chloride, chloroform, trichloroethylene	Past-practice 200-ZP-1 interim action (Freeman- Pollard 1996) ^(c)
Low-level burial grounds (Waste Management Areas 3 and 4)	Radioactive solid waste	Various chemical and radioactive wastes ^(d)	RCRA (Last and Bjornstad 1989)
616A crib (State- Approved Land Disposal Site)	Treated liquid effluent	Tritium	Active; WAC permitted (Barnett 2000a)
Environmental Restoration Disposal Facility	Excavated, contaminated soil and debris (potentially radioactive and/or hazardous)	None anticipated (double-lined facility)	Active (Weeks et al. 1996, Ford 1996)

⁽a) Sites with specific groundwater monitoring requirements and those that appear to have affected groundwater quality.

⁽b) Plan being revised.

⁽c) Groundwater monitored independently of groundwater project.

⁽d) Present in waste; not found in groundwater.

5.2 Conceptual Model

The most widespread hazardous chemical contaminants of concern in 200 West Area groundwater are, carbon tetrachloride, chloroform, and nitrate. Smaller plumes of chromium, fluoride, and trichloroethylene are also present. Iodine-129, technetium-99, tritium, and uranium are the most significant radionuclides in groundwater.

Contaminated effluent has reached the soil from cribs, trenches, tile fields, surface impoundments, and leaking tanks associated with T Plant, Reduction-Oxidation Plant, U Plant, and Plutonium Finishing Plant. Radionuclides with short half-lives decayed in the vadose zone, while nonradioactive constituents and longer-lived radionuclides were carried deeper. Some of these sorbed to sediment, some remained in the moisture in the vadose zone, and large quantities were carried into the groundwater.

The direction of groundwater flow beneath the southern portion of the 200 West Area is to the east. Groundwater flows to the northeast beneath the northern part of the area. In the past, waste disposal practices created groundwater mounds that caused some westward flow of contaminants. Contaminants moved outward from these mounds, contaminating a larger area in the saturated zone than in the vadose zone. These mounds are still present but are declining; the most recent information indicates that the westward flow has ceased. Interim remedial action systems, where groundwater is extracted, treated, and reinjected, locally perturb groundwater flow directions near the Plutonium Finishing Plant and east of U Plant.

The few shallow and deep well pairs indicate that the vertical flow gradient is downward in the 200 West Area. Contamination in the deeper parts of the unconfined aquifer appears to be considerably less than in the upper portion of the aquifer. It should be noted, however, that very few wells monitor the deeper portions of the aquifer, and at some locations certain contaminants are found at greater concentrations at depth.

Contaminant concentrations are expected to decrease with time because of dispersion, dilution, radioactive decay, remediation, and migration. There are no new sources of contamination, but concentrations of existing contaminants will vary because of plume movement and mobilization of vadose zone contamination.

5.3 Monitoring Program

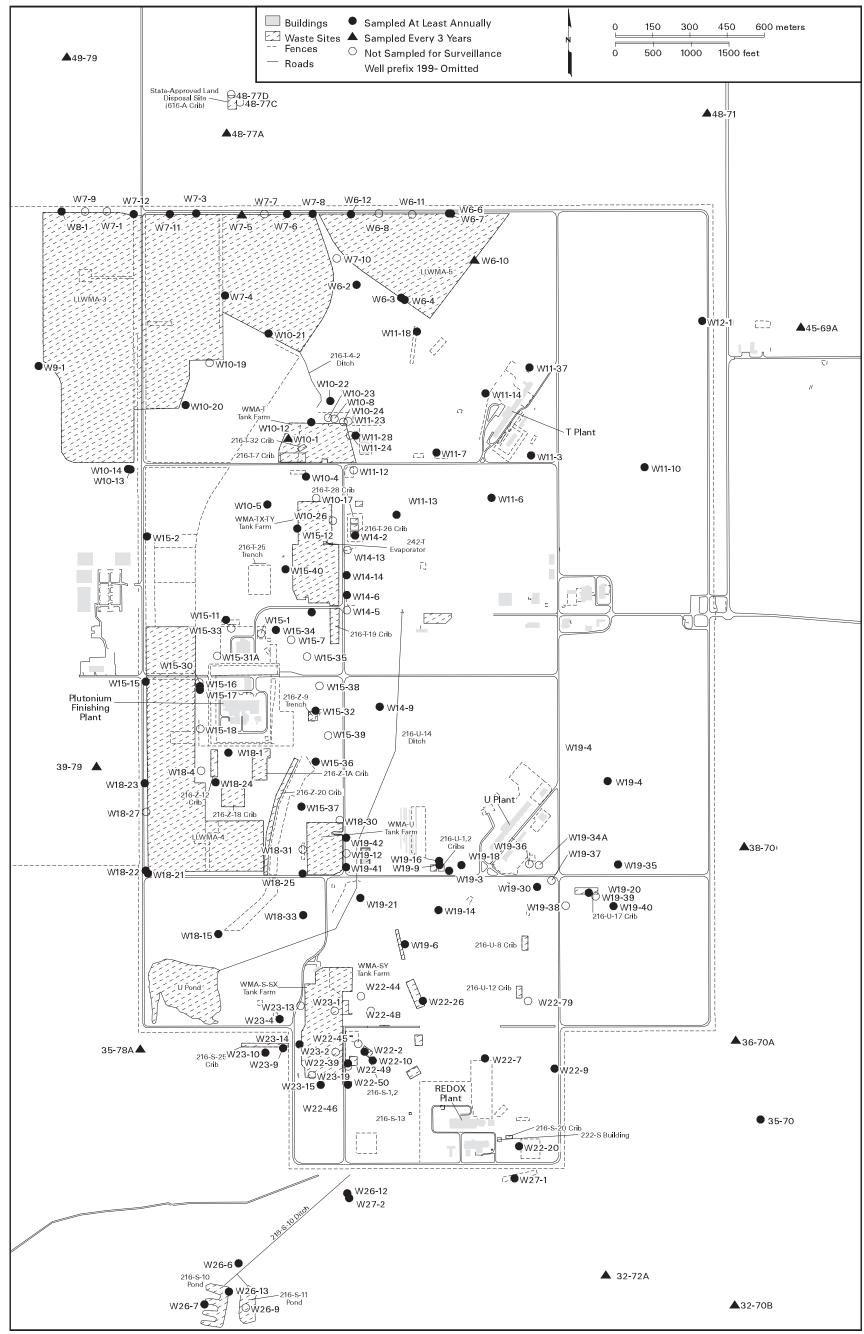
The primary objective of the surveillance monitoring program in the 200 West Area is to monitor the extent of plumes emanating from waste sites in the area. Most of the sites have ceased operation, and many wells monitoring the widespread plumes will be sampled annually since plume dimensions and concentrations continue to change significantly in this vicinity. Other wells will be sampled more frequently.

Another objective of surveillance monitoring is to deal with hazardous waste sites that ceased operation before 1985 and radioactive waste sites, which are not regulated by RCRA. Wells are monitored near the most significant sources to see if contaminants are declining as expected and to detect contaminants migrating from the vadose zone. These wells are monitored annually or more frequently.

The carbon tetrachloride network for monitoring the top of the aquifer was re-evaluated using a geostatistical analysis in fiscal year 2000. Carbon tetrachloride was chosen for this evaluation because it is the most widespread contaminant in the area and is a major driver for groundwater remediation. Many wells are monitored for carbon tetrachloride and the groundwater project felt that the geostatistical evaluation would improve the efficiency of the monitoring system. The analysis was carried out using the monitoring data from the fiscal year 1999 annual groundwater report (Hartman et al. 2000). Modeling the contaminant plume using variogram analysis was the first step in the process. The model then was used to produce stochastic simulations of the concentrations across the area. The stochastic simulations provided estimated concentration values (median value of 100 simulations) as well as measures of uncertainty at each location (interquartile range, etc.). The results of the analysis were used in a ranking procedure to determine which well provided the least valuable information. Wells close to other wells and wells in areas far from the plume core (as defined by the 2,000-µg/L contour) were rated less valuable, while remote wells and wells critical to defining the high concentration core were rated as higher value. After dropping the least valuable well from the well list, the simulations were iteratively re-run on the reduced well list, dropping the least valuable well each time. Thus, a ranking of wells was produced according to the value of the well in the monitoring network. Well rankings were adjusted using hydrogeologic interpretations. The analysis showed that 30 wells could be dropped from the carbon tetrachloride monitoring network with minimal degradation in the ability to map the plume. However, monitoring for carbon tetrachloride continues to be required for many of the wells because of CERCLA and RCRA monitoring agreements. The ranking was redone with a restriction on dropping wells with regulatory monitoring requirements and the results indicated that 16 wells could still be dropped.

Since the initiation of the geostatistical analysis, a large number of changes have been made in the carbon tetrachloride monitoring. These changes include wells going dry, new wells being drilled, and modifications to well and/or constituent lists for existing projects. To capture the most recent changes, the simulations are being rerun on the candidate fiscal year 2001 monitoring network. The results of the simulations will be used to reduce the monitoring network where possible. The modifications, however, are not reflected in this monitoring plan. They will be documented in project files through written change requests.

The locations of the monitoring wells in the 200 West Area are illustrated in Figure 5.1 (600 Area wells were shown in Figure 4.8). Wells and constituents are listed in Appendix A.



5.5

Figure 5.1. Groundwater Project Well Locations: 200 West Area

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6.0 200 East Area

For the purposes of this plan, "200 East Area" describes that portion of the Hanford Site within the 200 East Area fence line, those parts of the site downgradient from the area that shows impacts by contaminants originating in the 200 East Area, and those disposal facilities outside the fence line but associated with 200 East Area operations. Thus, B Pond, Gable Mountain Pond, and B/C cribs will be included in the following discussion. The 200 Areas Treated Effluent Disposal Facility also falls generally within this part of the site but is monitored under the specific requirements of its state waste discharge permit.

6.1 Background

6.1.1 Waste Sites, Discharges, and Groundwater Operable Units

Hundreds of waste sites have been identified in the 200 East Area, including radioactive and mixed waste storage tanks; low-level burial grounds; effluent disposal cribs, ditches, drains, and ponds; and various spills or other unplanned releases. Sites with specific monitoring requirements and those that appear to have affected groundwater quality are listed in Table 6.1. Additional information is provided in Hartman (2000). A number of facilities are regulated under RCRA because they were more recently active and contain, or contained, dangerous waste constituents.

The 200 Areas Treated Effluent Disposal Facility, located east of the 200 East Area proper, is the only active liquid disposal facility in the area. As mentioned above, this is monitored under a state waste discharge permit. The permitted discharge does not include radioactive or hazardous constituents. Low-Level Waste Management Areas 1 and 2 are burial grounds regulated under RCRA, which continue to receive radioactive solid waste. Three single-shell tank waste management areas, also regulated under RCRA, no longer actively receive waste but currently store mixed waste.

Groundwater in the northwestern part of the 200 East Area forms the 200-BP-5 Operable Unit, while the southeastern part of the site is in the 200-PO-1 Operable Unit. 200-BP-5 remediation is being performed under CERCLA regulations, while 200-PO-1 is being remediated under RCRA regulations (although requirements of both sets of regulations are considered in the cleanup process). Two groundwater extraction treatability tests were performed in the 200-BP-5 Operable Unit – the first near the 216-B-5 injection well and the second just north of the northwestern corner of the 200 East Area in an area of contamination originating in the BY cribs. There is no active groundwater remediation in the 200 East Area. The interim action recommended in the *Hanford Sitewide Groundwater Remediation Strategy* (DOE 1995a) is natural attenuation and decay of contaminant plumes. There is, however, no interim or final record of decision for the operable units in the 200 East Area.

Table 6.1. Selected Waste Sites in and Downgradient of the 200 East Area

Facility (period of use)	Waste Type	Constituents of Interest for Groundwater Monitoring	Type of Site (monitoring plan reference)	
	200 East Area Proper and Adjacent Facilities			
Single-shell tank farms (Waste Management Areas A-AX, B-BX-BY, C)	Radioactive/chemical slurries	Sodium hydroxide, sodium salts, radionu- clides, ferrocyanide	RCRA (Caggiano and Goodwin 1991, ^(a) Narbutovskih 2000)	
216-B-7A, -7B, -8 cribs	Supernatant from settling tanks	Sodium hydroxide, sodium salts, radionu- clides, ferrocyanide	Past-practice (200-BP-5 Operable Unit; TPA M-15-96-04)	
216-B-37 trench	Concentrated waste from tank bottoms	Sodium hydroxide, sodium salts, radionu- clides, ferrocyanide	Past-practice (200-BP-5 Operable Unit; TPA M-15-96-04)	
216-B-5 injection well (1945-46)	Hot cell drainage; supernatant from settling tanks	Strontium-90, cesium-137, plutonium	Past-practice (200-BP-5 Operable Unit; TPA M-15-96-04)	
BY cribs and trench (1954-55)	Uranium-recovery waste supernatant	Ferrocyanide, radionuclides	Past-practice (200-BP-5 Operable Unit; TPA M-15-96-04)	
216-B-63 trench (1970-92)	Steam condensate	Sulfuric acid, sodium hydroxide, radionuclides	RCRA (Sweeney 1995a)	
Plutonium-Uranium Extraction Plant waste- disposal cribs	Process distillate	Radionuclides (especially tritium, iodine-129, strontium-90), nitrate	216-A-10, -36B, -37-1: RCRA (Lindberg 1997); others: past-practice (200-PO-1 Operable Unit; DOE 1997d)	
216-A-29 ditch	Plutonium-Uranium Extraction Plant chemi- cal waste	Sodium hydroxide, sulfuric acid	RCRA (Sweeney 1999)	
216-B-3 pond (B Pond)	B Plant steam condensate and chemical waste; Plutonium-Uranium Extraction Plant chemical waste	Tritium, aluminum nitrate, potassium hydroxide, nitric acid, sulfuric acid	RCRA (Sweeney 1995b) ^(a)	
200 Areas Treated Effluent-Disposal Facility	Treated liquid effluent from 200 Areas	Trihalomethane	Active; WAC permitted (Barnett 2000b) ^(b)	
Low-Level Burial Grounds (Waste Management Areas 1 and 2)	Radioactive solid waste	Various chemical and radioactive wastes ^(c)	RCRA (Last and Bjornstad 1989)	

Table 6.1. (contd)

Facility (period of use)	Waste Type	Constituents of Interest for Groundwater Monitoring	Type of Site (monitoring plan reference)
Liquid Effluent Retention Facility	242-A evaporator process condensate	Ammonium, acetone, aluminum, 1-butanol, 2-butanone, tritium, strontium-90, ruthenium-106, cesium-137 ^(c)	RCRA (Schmid 1990)
600 Area Facilities			
Gable Mountain Pond (1957-87)	200 East Area liquid wastes	Strontium-90, cesium-137, ruthenium-106	Inactive (this document)
BC Cribs and Trenches (1956-58)	Uranium recovery waste supernatant	Ferrocyanide, radionuclides	Past-practice (this document)
(a) Plan being revised.(b) Groundwater monitored independently of groundwater project.(c) Present in waste; not found in groundwater.			

6.1.2 Groundwater Monitoring Requirements and History

Groundwater monitoring has been conducted in the 200 East Area since the 1940s. Very few monitoring wells existed in the early decades but more were installed in the 1970s and monitored per DOE requirements. Approximately 100 new wells were installed when RCRA monitoring began in the late 1980s. CERCLA investigations in the 1990s resulted in the installation of several wells but relied primarily on data from existing groundwater monitoring networks and additional wells installed in support of RCRA.

The *Atomic Energy Act of 1954* and DOE Order 5400.1 require monitoring to identify and track contaminant plumes. This document serves as the monitoring plan for surveillance monitoring per DOE orders.

6.2 Conceptual Model

The most widespread groundwater contaminants of concern originating from the 200 East Area are iodine-129, nitrate, and tritium. These contaminants extend east and southeast of the 200 East Area to the Columbia River and northwest to the gap between Gable Mountain and Gable Butte. A significant plume of technetium-99 at levels above the drinking water standards extends northwest from the 200 East Area fence line toward the gap between Gable Mountain and Gable Butte. This plume area also contains low levels of cobalt-60 and cyanide. Arsenic is found at levels above drinking water standards in the eastern part of the 200 East Area. Groundwater is locally contaminated with strontium-90 at high levels near Gable Mountain Pond (decommissioned) and at low levels near cribs south of the Plutonium-Uranium

Extraction Plant. Contamination with cesium-137, plutonium, and strontium-90 is found in the immediate vicinity of the 216-B-5 injection well. Localized uranium and chromium contamination is also found.

The most extensive contaminant plumes are attributable predominantly to liquid discharges to cribs, with some contribution from ponds, ditches, and other sources. Most pond discharge, however, was more dilute and did not contribute to the highest levels of contamination. The ponds, particularly 216-B-3 Pond (B Pond), did have a large influence on contaminant migration because the large amounts of water that went to the ponds affected flow directions. Groundwater mounding at B Pond remains evident, though the mound is declining since discharge to the pond ceased. Contamination from tank leaks, unplanned releases, and specific retention trenches appears to have produced groundwater contamination of limited extent, though considerable inventory may remain in the vadose zone. No groundwater impact from low-level waste burial grounds in the 200 East Area has been identified.

Contaminant levels are declining through much of the area. Many short-lived radionuclides detected in the past, such as cobalt-60 and ruthenium-106, are no longer detected or are detected at much lower concentrations. Tritium concentrations near the source areas are declining because of termination of discharge and the subsequent dispersion and decay within the plume. Residual contamination in the vadose zone at many of the sources continues to drain into the groundwater. The amount of transport to groundwater may decline with time. Some contaminants that have been retarded by sorption to sediment or that never reached groundwater because of limited discharge volumes (i.e., specific retention trenches) may break through to the water table, and concentrations then could increase. In addition, any uncontrolled discharge, such as leaks from water lines, may enhance transport of contaminants to the groundwater from the vadose zone. Tritium concentrations north of the 300 Area near the Columbia River continue to increase, reflecting the continued spread of contamination at the downgradient plume boundaries. The contamination is predicted not to spread southward beyond the site boundary (Hartman and Dresel 1998).

Several waste streams discharged at the site included chemical complexants along with other radioactive and hazardous constituents. Complexation is known to have enhanced the mobility of some constituents such as cobalt-60. Cyanide, a complexant for cobalt-60, is monitored in the northern part of the area and adjacent 600 Area. Cobalt-60 concentrations are declining because of its short half-life, but other constituents such as americium, plutonium, and neptunium may also have increased mobility. The data on these constituents and on complexant concentrations are sparse. More work is needed to evaluate the potential for complexant-enhanced mobility of radionuclides before including them in routine monitoring.

Vertical migration of contaminants to deeper parts of the aquifer or deeper aquifers may have occurred through several mechanisms. Significant groundwater mounds developed at a number of facilities. By far, the greatest mounding occurred at B Pond, where monitoring evidence indicates there was some movement of contamination down to the upper basalt-confined aquifer. This mounding produced vertical gradients to transport contamination downward in the sedimentary sequence. Poorly sealed wells may have produced conduits, thus enhancing vertical migration.

Additional vertical migration probably occurred through the discharge of high-density, high-salt wastes. This dense material migrated down through the aquifer. Few monitoring data are available on the lower parts of the unconfined aquifer because of the paucity of monitoring wells. Dense waste could have moved down-dip along the top of basalt. The unconsolidated sediment is thin in the area where high-density waste was discharged, increasing the potential for the waste to reach the basalt. Further evaluation of areas where this type of vertical migration may have occurred is needed to determine additional monitoring needs.

A third mechanism for vertical migration is the intersection of the water table by confining layers in the suprabasalt sediment. The lower Ringold mud intersects the water table downgradient of B Pond and dips approximately to the south. This serves to induce downward flow to the sediment below the confining mud. The lower part of the Ringold sediment, thus, forms a confined aquifer in this area. Although relatively few wells are completed below the lower Ringold mud, this interval is monitored where wells are available. Several wells near B Pond and the 200 Area Treated Effluent Disposal Facility are completed below the lower mud.

6.3 Monitoring Program

The surveillance monitoring program in the 200 East Area has been designed to meet several objectives and to complement the RCRA monitoring networks. The first objective is to monitor the extent of plumes emanating from 200 East Area waste sites. Most of these sites have ceased operations and, thus, it is expected that the monitoring network will be suitable to track the rate of dissipation and attenuation of the plumes. A combination of geostatistical assessment and site knowledge was used to develop the plume-monitoring system. The planned network has been assessed with a geostatistical model of the plumes, and the network was determined to provide sufficient data for plume tracking. The model used stochastic simulation to provide estimates of concentrations throughout the plume area and a measure of the uncertainties associated with the plume model. These uncertainties were used to rank wells according to their importance to the statistical model. Wells close to other wells and wells in areas where concentrations do not vary greatly received low rankings, while remote wells and wells where concentrations vary considerably over short distances were ranked higher. These rankings were then used by groundwater project scientists to choose wells to delete from the network. The geostatistical model was then used to evaluate the proposed network to ensure that similar results could be obtained with fewer monitoring wells.

The large contaminant plumes do not change rapidly. Therefore, many of the wells are sampled at one to three year intervals rather than more frequently.

Three bands of guard wells will be monitored annually for a longer list of constituents to ensure that the nature of contamination found downgradient of the operational and waste-disposal areas has been sufficiently characterized. These bands are shown in Figure 6.1; the wells are listed in Table 6.2. One band is located in the gap between Gable Mountain and Gable Butte and serves to detect contaminant movement to the north. The second band is located to the southeast of the 200 East Area and detects contamination moving into the southern and eastern parts of the site. The third band is along the Columbia River to provide assurance that offsite impacts are identified. In addition to the known contaminants,

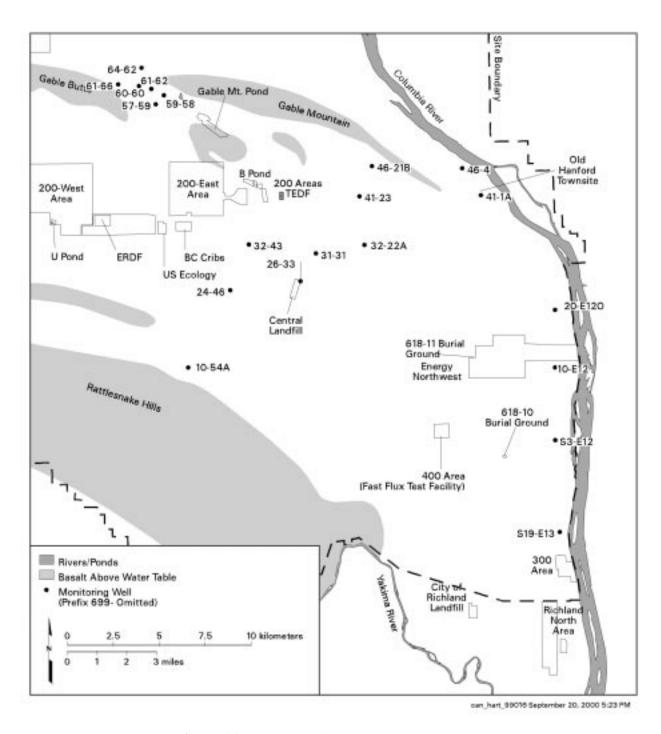


Figure 6.1. Locations of 200 East Area Guard Wells

Table 6.2. 200 East Area Guard Wells

Gap	Southeast	River
699-57-59	699-10-54A ^(a)	699-10-E12
699-59-58	699-24-46	699-20-E12O
699-60-60	699-26-33	699-41-1A
699-61-62	699-31-31	699-46-4
699-61-66	699-32-22A	699-S3-E12
699-64-62	699-32-43	699-S19-E13 ^(b)
	699-41-23 ^(a)	
	699-46-21B ^(a)	

Constituent List: Inductively coupled-plasma metals; anions; gross alpha, beta, and gamma; strontium-90; technetium-99; tritium; total organic halides; total organic carbon; and alkalinity.

- (a) Reduced list Sample tritium, alpha, beta, anions annually; Full list Sample every 3 years.
- (b) Also monitors southern portion of the Hanford Site.

wells in these bands will be monitored for inductively coupled-plasma metals, anions, gross alpha, gross beta, gamma, strontium-90, technetium-99, tritium, total organic halides, total organic carbon, and alkalinity.

The monitoring network is also designed to complement the RCRA detection and assessment monitoring of contaminant sources. The RCRA-monitoring networks only address sources, containing hazard-ous constituents, that were operational after 1985. RCRA and past-practice source monitoring serves to ensure that concentrations of groundwater contaminants are declining near the most significant sources and to detect the breakthrough of new contamination from the vadose zone. The source monitoring places a high priority on the potentially largest contaminant sources and on areas near the 200-East Area boundary. This helps ensure that any new contamination will be detected before it moves out of the operational area.

Locations of monitoring wells for the 200 East Area are illustrated in Figure 6.2 (wells in the 600 Area were shown in Figure 4.8). Wells and constituents are listed in Appendix A.

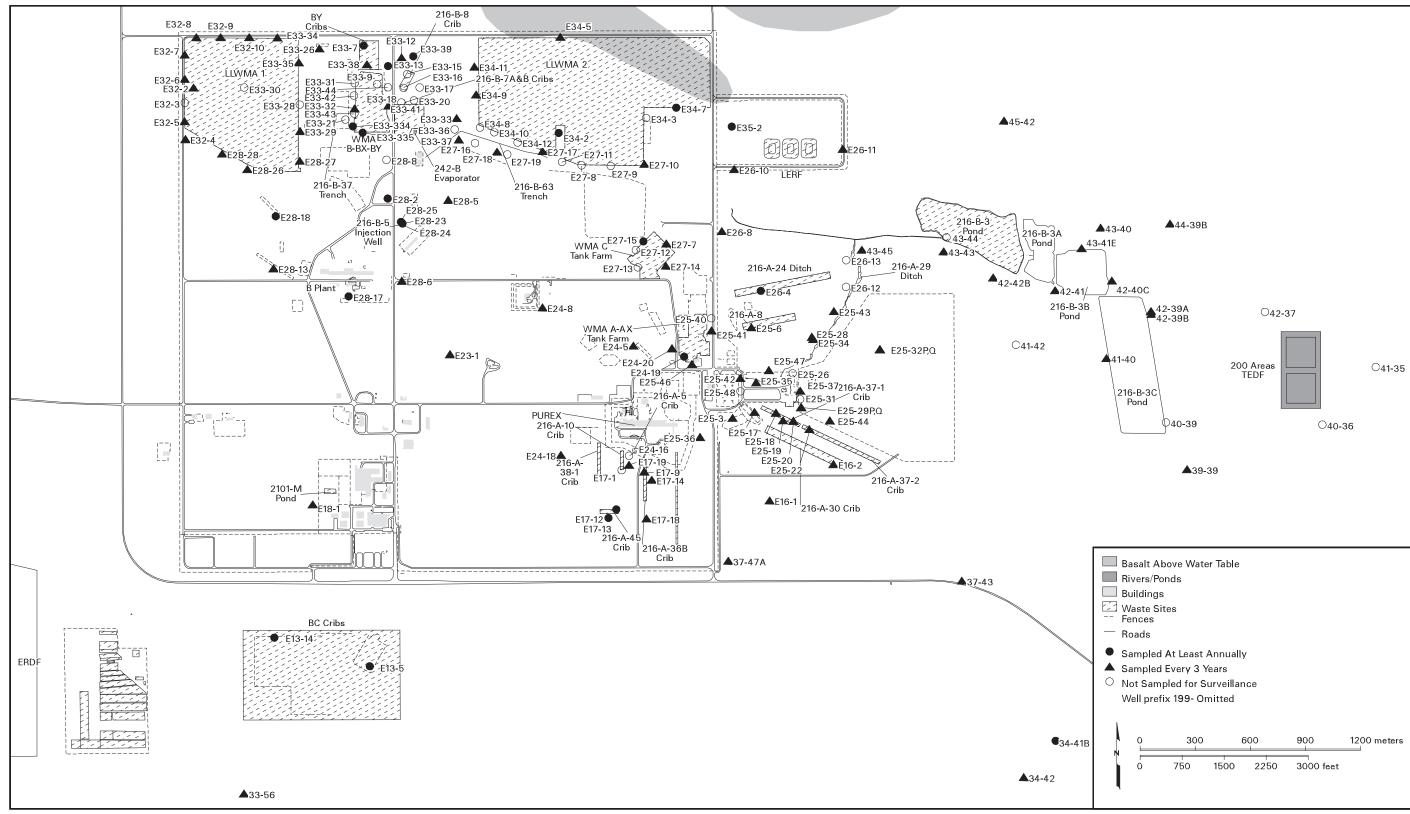


Figure 6.2. Groundwater Project Well Locations: 200 East Area

7.0 400 Area

7.1 Background

This section covers activities in the 400 Area, the location of the Fast Flux Test Facility, a liquid sodium-cooled reactor. The reactor is on standby, pending a restart decision for the production of medical isotopes and tritium.

Primary local groundwater monitoring includes the area around the 4608 B/C ponds (also called the 400 Area process ponds), which receive wastewater effluent. The water supply for the 400 Area, including the drinking water supply, is also monitored by sampling wells completed in the unconfined aquifer system.

7.1.1 Waste Sites

The 400 Area process ponds are located north of the 400 Area perimeter fence and are unlined infiltration ponds that receive wastewater from the 400 Area facilities (Figure 7.1). The waste stream consists primarily of cooling water and intermittent small contributors such as sinks and drains.

7.1.2 Groundwater Monitoring Requirements and History

The 400 Area process ponds are monitored in accordance with State Waste Discharge Permit ST4501, issued on August 1, 1996 and modified on February 10, 1998. This integrated groundwater monitoring plan provides requirements for sampling activities and quality assurance/quality controls to ensure that the data needs of various users are satisfied. The primary objective of groundwater monitoring at this facility is to ensure that wastewater entering the ponds meets acceptable standards and does not adversely affect local groundwater quality. The monitoring network includes two downgradient wells (699-2-6A and 699-2-7) and an upgradient well (699-8-17), shown in Figure 4.8. Constituents analyzed in quarterly groundwater samples, as specified by the discharge permit, include unfiltered metals (cadmium, chromium, lead, mercury, and manganese), pH, sulfate, and total organic carbon. In addition, the wells are co-sampled for surveillance monitoring for other constituents, as indicated in Appendix A.

Nitrate is the only contaminant that has been consistently identified at concentrations above regulatory limits in the groundwater monitoring network for the 400 Area process ponds, where it has been monitored in well 699-2-7 since 1986. This is attributed to a sanitary sewage lagoon formerly located immediately west and upgradient of the ponds and later to a drainfield associated with septic tanks located southwest of the ponds. Disposal to the lagoon was discontinued in 1983 or 1984 and to the drainfield in April 1997; thus, groundwater contamination from these sources is expected to diminish with time. Nitrate concentration levels in well 699-2-6A have been monitored since 1998. The levels are relatively low, apparently due to the low nitrate content of the effluent disposed to the ponds.

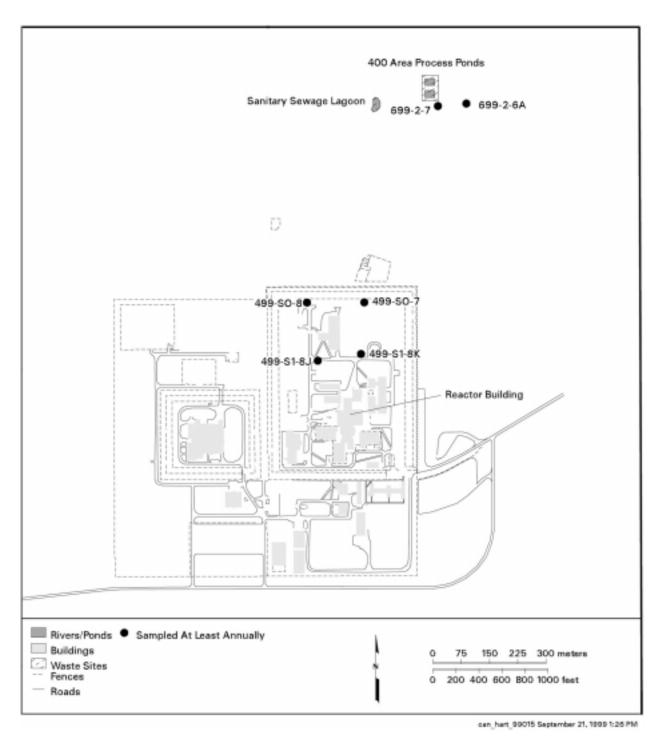


Figure 7.1. Groundwater Project Well Locations: 400 Area

The primary compliance issue for groundwater monitoring related to the 400 Area water supply is related to tritium. Wells 499-S0-7 and 499-S0-8, the original water-supply wells, were completed near the top of the unconfined aquifer and have been monitored since 1972. When tritium contamination was detected in the water supply, an additional well (499-S1-8J) was drilled in the lower unconfined aquifer in 1985. Tritium levels of water samples collected from well 499-S1-8J are well below the interim drinking water standard of 20,000 pCi/L, and this well is currently used as the primary water-supply well. Wells 499-S0-7 and 499-S0-8 are still used for backup supply and emergency uses, but based on DOE direction may not be used as a potable water supply when the drinking water standard is exceeded.

7.2 Conceptual Model

Water-level contours indicate that groundwater generally flows from west to east across the 400 Area. In addition, tritium and nitrate plumes, which originate in the 200 East Area, indicate that groundwater flows toward the east to southeast. The tritium plume is detected in the 400 Area water-supply wells, as discussed in Section 7.1.2. Tritium levels are lower in the vicinity of the 400 Area process ponds as a result of dilution effects. However, nitrate levels are currently elevated in the vicinity of the process ponds, apparently from the former disposal of sanitary sewage to a nearby lagoon. Since discharge of process water to the ponds is monitored under a state waste discharge permit, migration of contaminants from the ponds to the groundwater is not expected to be significant.

7.3 Monitoring Program

Monitoring well locations in the 400 Areas are presented in Figure 7.1. Upgradient well 699-8-17 was shown in Figure 4.8. Constituents monitored and sampling frequencies are presented in Appendix A.

8.0 300 and Richland North Areas

The 300 and Richland North Areas include the southern portion of the Hanford Site and adjacent area to the south between the Yakima and Columbia Rivers. The 300 Area is located along the Columbia River in the southeastern portion of the Hanford Site. The Richland North Area, though not formally defined, includes the former 1100 Area and the 3000 Area, part of the 600 Area adjacent to the 300 Area, and parts of nearby Richland. Figure 8.1 shows the locations of these two areas.

8.1 Background

The 300 Area has been used for research and development and nuclear fuel fabrication associated with uranium fuel elements for nuclear reactors. The Richland North Area consists of a variety of both onsite and offsite land uses, including municipal, industrial, and agricultural. Municipal and industrial facilities and agricultural activities in the Richland North Area influence groundwater. Offsite facilities of particular interest with respect to groundwater include the city of Richland's North Well Field and recharge basins, the ORV Park, Siemens Power Corporation, Richland Landfill, Lamb-Weston Richland Plant, Interstate Nuclear Services, and Allied Technology Group. Offsite agricultural irrigation influences groundwater over a wide area in the Richland North Area.

8.1.1 Waste Sites, Discharges, and Groundwater Operable Units

In the 300 Area, inactive waste sites known to have received liquid waste containing uranium and other known or suspected contaminants include the 316-5 process trenches and the 316-1 and 316-2 process ponds. These are the primary sites affecting groundwater contamination. Other sites that received wastes include sanitary septic tanks, trenches, and tile fields; ash pits; filter backwash ponds; and a number of burial grounds. The 300 Area also contained underground tanks for storing gasoline and diesel fuels.

The 316-5 process trenches require groundwater monitoring to meet RCRA requirements because the trenches are regulated as dangerous waste surface impoundments. The process trenches were modified as part of an expedited response action in 1991, and discharges to the trenches ceased in late 1994. The 316-1 and 316-2 process ponds, monitored to meet CERCLA requirements, received uranium-contaminated wastewater until 1975 when the process trenches began receiving the wastewater. The storage tanks were monitored under the state's underground storage tank program in the early 1990s, but monitoring is no longer required by the state. Groundwater in the 300 Area is part of the 300-FF-5 Operable Unit.

Waste sites in the Richland North Area include the inactive Horn Rapids Landfill in the 600 Area and a number of disposal pits and underground storage tanks in the former 1100 Area. Groundwater associated with these waste sites is monitored to meet CERCLA requirements. The 1100-EM-1 Operable Unit record of decision(1993) required groundwater monitoring at a point of compliance downgradient from the inactive Horn Rapids Landfill. There are no DOE waste sites requiring RCRA groundwater

monitoring in the Richland North Area. Table 8.1 lists the waste sites in the 300 and Richland North Areas that have specific monitoring requirements and those sites that affected groundwater quality.

8.1.2 Offsite Sources

Probable sources of groundwater contamination that originated from the Richland North Area off the Hanford Site include Siemens Power Corporation, agricultural irrigation, and Lamb-Weston Richland Plant. Siemens Power Corporation is located adjacent to the Hanford Site boundary southwest of the Horn Rapids Landfill, and a surface impoundment system at the site contributed to solvent and nitrate contamination in groundwater. Fertilizers applied to the agricultural fields upgradient (south) of Siemens Power Corporation and potato-processing waste from the Lamb-Weston Richland Plant are probable sources of nitrate. The Richland Landfill, Interstate Nuclear Services, and Allied Technology Group are not known to contribute to groundwater contamination on the Hanford Site.

The city of Richland's North Well Field and recharge basins, located in the south-central part of the Richland North Area, are the primary influence on changes to groundwater elevation in the area. The well field serves as a secondary water supply for the city of Richland, and the basins recharge the unconfined aquifer with Columbia River water. The net recharge causes a groundwater mound to form in this area and decreases nitrate levels in groundwater to less than ambient.

Irrigation applied to agricultural fields contributes to groundwater recharge during the growing season. As a result, this contributes to groundwater flow to the northeast, east, and southeast.

8.1.3 Groundwater Monitoring Requirements and History

Extensive groundwater monitoring has been conducted in the 300 Area as far back as 1975, when the 316-5 process trenches replaced the 316-1 and 316-2 process ponds as the main facility for disposal of uranium-contaminated wastewater. The earliest major study on groundwater contamination in the 300 Area was in 1977 (Lindberg and Bond 1979). A site-specific program of groundwater monitoring of the 300 Area has been conducted since 1977. In 1985, interim status groundwater monitoring of the process trenches was initiated under RCRA, which required additional wells to be installed (Schalla et al. 1988). The RCRA program went into final-status groundwater monitoring in 1996 (Lindberg et al. 1995). In response to the Tri-Party Agreement (Ecology et al. 1998), CERCLA activities were initiated in the early 1990s and included additional groundwater monitoring. An expedited response action was implemented in 1991 to remove sources of contamination and resulted in lower contaminant concentrations in groundwater downgradient from the process trenches. An interim remedial action required continued groundwater monitoring of contaminants in the 300 Area (ROD 1996b).

Groundwater well installation and monitoring began in the 1100 Area in 1988 after a limited groundwater-sampling effort in 1986 revealed low levels of contaminants (DOE 1990). A study was conducted in 1989-1992 and included well installation and groundwater monitoring to determine the nature and extent of groundwater contamination in the 1100 Area (DOE 1992). In 1993, continued and expanded groundwater monitoring in the 1100 Area was required (ROD 1993). In response to the record of decision, additional well installation and monitoring were implemented at the inactive Horn Rapids

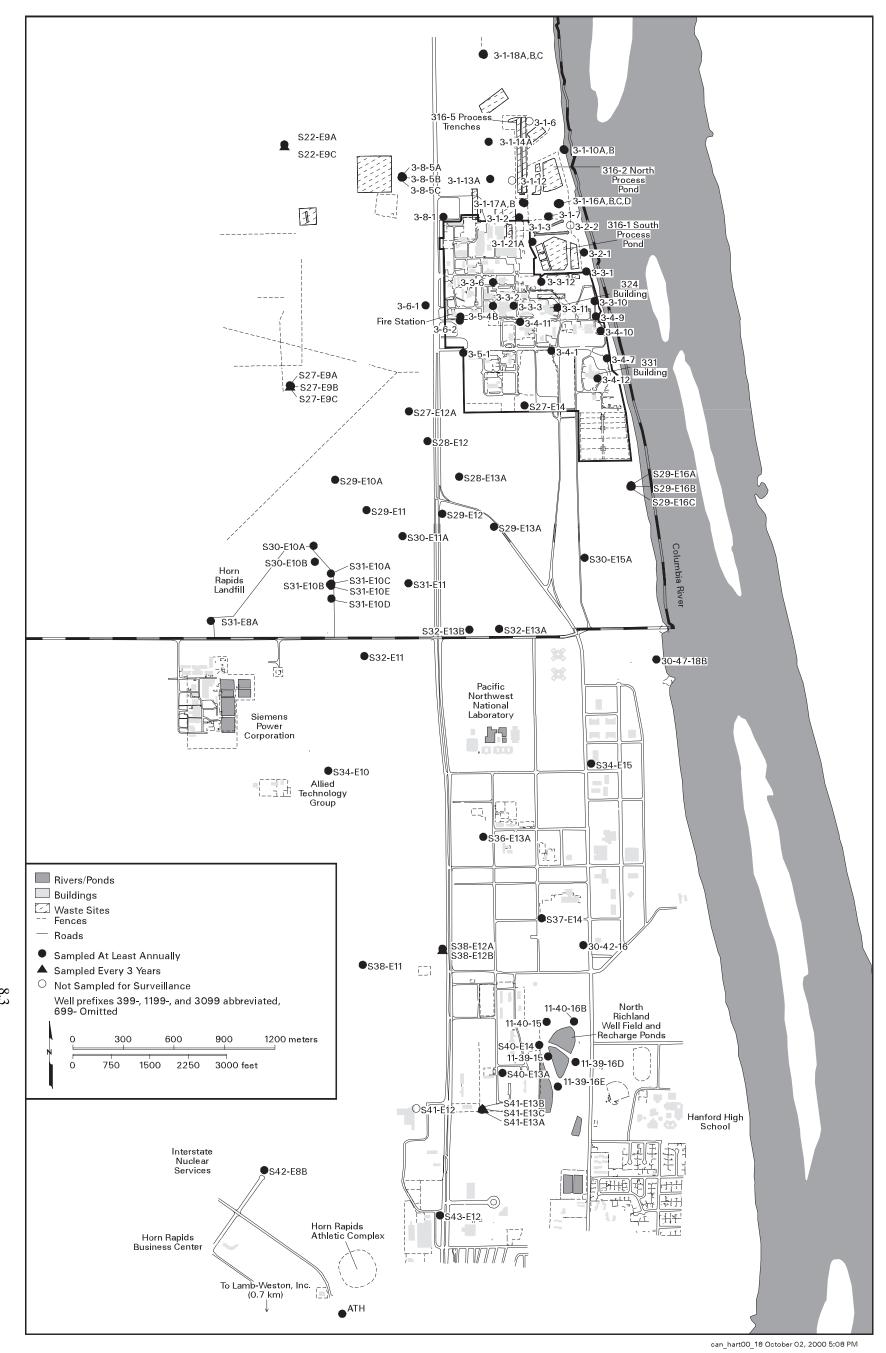


Figure 8.1. Groundwater Project Well Locations: 300 and Richland North Areas

Table 8.1. Selected Waste Sites in the 300 and Richland North Areas

Escilita (social of use)	We ste Tour	Constituents of Interest for Groundwater	Type of Site (monitoring plan
Facility (period of use)	Waste Type	Monitoring	reference)
	300 Area		
316-5 process trenches (1975-94)	Variety of chemical and uranium wastes	Uranium, trichloro- ethylene, cis-1,2- dichloroethylene, metals	RCRA (Lindberg et al. 1995)
316-1 (south) and 316-2 (north) process ponds (1940s-75)	Variety of chemical and uranium wastes	Uranium, trichloro- ethylene, cis-1,2- dichloroethylene	Past-practice (300-FF-1 and -5 Operable Units, ROD 1996b, DOE 1996a)
Richland North Area			
Horn Rapids Landfill (1950s-70)	Office and construction wastes, asbestos, sewage sludge, fly ash	Trichloroethylene, break- down products of trichloroethylene (vinyl chloride, 1,1-dichloro- ethylene), chromium, technetium-99, nitrate	Past-practice (1100-EM-1 Opera- ble Unit, ROD 1993, DOE 1995b)
Siemens Power Corporation process lagoons (offsite) (1971-present)	Ammonia, fluoride, nitrate, radionuclides (primarily uranium)	Trichloroethylene, nitrate	Active RCRA; Siemens (1996) ^(a)
Lamb-Weston (offsite)	Potato-processing wastes	Nitrate	Active
Agriculture (offsite)	Fertilizers	Nitrate	Active
(a) Groundwater monitored independently of groundwater project.			

Landfill (DOE 1995b). The record of decision required the monitoring of trichloroethylene in ground-water downgradient of the inactive Horn Rapids Landfill. The record of decision also required monitoring of trichloroethylene breakdown products and nitrate. The monitoring of nitrate was needed because its concentrations were above the maximum contaminant levels for nitrate.

The surface impoundment system at Siemens Power Corporation consists of six lagoons, which are regulated under the Revised Code of Washington, Title 70, Chapter 105 (Siemens Power Corporation 1997). The lagoons no longer receive liquid waste, however the lagoons still contain liquid wastes from disposal in the past. Siemens Power Corporation has monitored groundwater at their facility since 1994 to meet the requirements of RCRA interim status facilities.

8.2 Conceptual Model

The most widespread contaminants of concern in the 300 and Richland North Areas are tritium and nitrate, but these have sources outside the area. Waste sites within the 300 Area have contaminated

groundwater with uranium, cis-1,2-dichloroethylene, and trichloroethylene. The movement and distribution of these contaminants in groundwater are heavily influenced by Columbia River stage, a recharge mound at the Richland North Well Field, and agricultural irrigation practices.

The tritium plume is derived from past wastewater disposal in the 200 East Area and represents the southern margin of the sitewide plume that is encroaching into the 300 Area. Tritium migrates across the northeastern portion of the 300 Area from the north and enters the Columbia River. Tritium levels have generally been steady with time in and north of the 300 Area in recent years. The southward migration of tritium is limited to the 300 Area because of the following factors:

- Groundwater is recharged by the Yakima River, and flows generally from southwest to northeast, and discharges to the Columbia River.
- Recharge from agricultural irrigation and an unlined artificial pond at the ORV Park between the Yakima River and the former 1100 Area contributes to eastward groundwater flow.
- Net recharge at the city of Richland's North Well Field has resulted in a groundwater mound that directs groundwater flow outward, including a component to the north.

In the 300 Area, wastewater effluent, containing uranium and chlorinated solvent compounds, percolated through the soils from leaking process trenches and ponds for approximately 50 years. These constituents were driven down through the soils in the vadose zone beneath the waste sites by subsequent effluent discharges and natural recharge. As the constituents were carried downward, some were sorbed to sediments and trapped in soil moisture and some reached groundwater. Uranium in groundwater migrates toward and enters the Columbia River.

Uranium concentrations in groundwater fluctuate indirectly in response to river-stage changes. As the river stage rises, groundwater near the river rises into a portion of the vadose zone. As a result, uranium is desorbed from the sediments and mobilized, increasing the uranium concentrations in groundwater. As the groundwater levels drop, uranium concentrations decrease because the thickness of the saturated sediments from which uranium desorbs decreases.

Chlorinated solvent compounds are generally found in the deeper portion of the unconfined aquifer beneath the process trenches and ponds. They have been found in the upper portion of the unconfined aquifer for brief periods (e.g., tetrachloroethylene in 1998). Two conceptual model hypotheses have been suggested for the deeper occurrences. One hypothesis is that dissolved chlorinated compounds in groundwater were transported by a downward vertical hydraulic gradient created when discharged effluent to the ponds and trenches recharged the aquifer, causing groundwater levels to rise. The second hypothesis is that an immiscible phase that is denser than water was driven to the bottom of the unconfined aquifer by density and rested on top of the silty clay unit. A portion of the dense phase would then dissolve into the aqueous phase.

A nitrate plume is migrating toward the 300 Area from the southwest toward the Columbia River. This area also contains a trichloroethylene plume. Nitrate contamination is the result of offsite industrial

and agricultural uses. Wastewater effluent containing ammonia was discharged in the past to lagoons at Siemens Power Corporation. Effluent has apparently leaked to the underlying soils from the lagoons, and some of the ammonia reached groundwater. Under aerobic conditions, the ammonia degrades relatively quickly to nitrate, which is highly mobile in groundwater. In agricultural areas to the southwest, fertilizers containing nitrate are applied during the growing season. As irrigation is applied, the dissolved nitrate is carried downward through the soils and is taken up by crops in the root zone. However, some of the nitrate is carried downward below the root zone by recharge of excess irrigation and reaches groundwater.

Trichlorothylene contamination is suspected to be the result of offsite industrial solvent use at Siemens Power Corporation. Solvents were used during installation, cleaning, and repairing of lagoon liners over a 10-year period between 1978 and 1988. Excess solvents contacted the soils by spillage and were driven downward into the vadose zone and reached groundwater, which is very shallow in this area. On reaching groundwater, trichloroethylene is very mobile and formed a localized plume that migrated downgradient to the northeast across the Horn Rapids Landfill. The highest concentrations were found near Siemens Power Corporation and the Horn Rapids Landfill. Trichloroethylene concentrations were measured as high as $420~\mu\text{g/L}$ in the late 1980s, but decreased to less than $10~\mu\text{g/L}$ by the late 1990s. One hypothesis has been suggested that natural attenuation may have reduced the mass of the trichloroethylene in groundwater. Natural attenuation in groundwater can occur by volatilization through passive pumping and biodegradation. Measurable trichloroethylene concentrations were observed in soil gas in the vicinity of the inactive Horn Rapids Landfill (Evans 1989). In fiscal year 1999, only two wells showed trichloroethylene concentrations above the maximum contaminant level. The wells are located immediately downgradient of the Horn Rapids Landfill.

8.3 Groundwater Monitoring Network

Wells, constituents, and frequency of monitoring are listed in Appendix A. Well locations were shown in Figure 8.1. Geostatistical assessment and current knowledge of the 300 and Richland North Areas were used in developing the monitoring network. Geostatistical assessment was used only in developing the monitoring network for the primary constituents of concern (see Chapter 6).

Uranium is the primary constituent of concern in groundwater in the 300 Area. Strontium-90, cis-1,2-dichloroethylene, and trichloroethylene are elevated locally. Trichloroethylene and nitrate from offsite sources contaminate groundwater in the Richland North Area. Special considerations in these areas include tracking the movement of the leading edges of tritium and other plumes that are near the city of Richland, monitoring effects of river-stage changes, and monitoring wells downgradient from potential offsite contaminant sources.

One objective is to monitor the extent of groundwater contamination in the 300 and Richland North Areas to ensure that contaminants have not migrated offsite and have not affected wells in the city of Richland. This requires intensive monitoring near the leading edges of the plumes, in areas along the site boundary, and in areas where concentrations are low. Monitoring in areas where levels are low provides a baseline from which to determine concentration changes and, thus, early detection of offsite migration. In fiscal year 2001, monitoring will be increased to include more wells completed in the deeper part of the

unconfined aquifer. These wells will be sampled every three years for low-level tritium analyses. An enrichment technique is used to measure tritium at lower detection limits (\sim 10 pCi/L) than provided by the standard method.

Another objective of monitoring the Richland North Area is to define plumes that have migrated onto the site from offsite sources. This monitoring is needed to show effects to onsite groundwater and to show that groundwater contamination attributed to these plumes is not derived from onsite waste sites.

9.0 600 Area Non-Operational Monitoring Activities

9.1 Background

The 600 Area includes those parts of the Hanford Site not specifically included within the boundaries of the operational areas, though many of the 600 Area wells serve to monitor large contaminant plumes with their sources in the operational areas. Those wells are discussed in previous sections, especially Chapter 6, 200 East Area. This section largely addresses those parts of the 600 Area of the Hanford Site not included in the monitoring activities associated with the operational areas discussed in the other sections. The groundwater project monitors several waste sites located in the 600 Area, and monitors wells outside of Hanford influences to provide a basis for defining background groundwater chemistry. In addition, monitoring of chemistry and hydraulic head data is conducted within confined aquifers. Specifically, the region addressed in this section is that portion of the 600 Area west of the 200 West, east and north of the Columbia River, and several facilities not covered in other sections of this plan.

9.1.1 Waste Sites

The Central Landfill, 618-10 and 618-11 burial grounds, and 316-4 crib are facilities located in the 600 Area that are not included in other sections of this plan and are discussed below. The Gable Mountain Pond and BC Cribs were included in Chapter 6.0 because they were associated with 200 East Area operations.

Agricultural activities in the area west of the Hanford Site contribute nitrate to the western portion of the 600 Area. Similar impacts of agriculture are recognized in the 600 Area north and east of the Columbia River.

9.1.1.1 Central Landfill

The Central Landfill is located approximately 5.5 kilometers southeast of the 200 East Area and consists of the Solid Waste Landfill and the Nonradioactive Dangerous Waste Landfill, which are currently monitored separately under different regulations (Table 9.1).

9.1.1.2 618-10 and 618-11 Burial Grounds and 316-4 Crib

The 618-10 burial ground and adjacent 316-4 crib are located southeast of the 400 Area. The 618-11 burial ground is located immediately northwest of Energy Northwest (see Figure 4.8). The burial grounds operated from 1954 to 1963 and received a variety of low- to high-activity radioactive waste, mostly composed of fission products with some plutonium-contaminated material (DOE 1996b). The waste was disposed in caissons and trenches and may have included liquid and solid waste forms.

Table 9.1. Selected Waste Sites in the 600 Area

Facility (period of use)	Waste Type	Constituents of Interest for Groundwater Monitoring	Type of Site (monitoring plan reference)
618-10 and -11 burial grounds and 316-4 crib (1948-62)	Low- to high-activity radioactive waste	Tritium, nitrate, hexone, organic wastes	Past-practice (300-FF-2 and 300-FF-5 Operable Units; DOE 1996a ^(a)
Solid Waste Landfill	Solid waste, sewage, garage wash water	Organics	WAC permitted (Hodges 1993) ^(b)
Nonradioactive Dangerous Waste Landfill	Asbestos, laboratory wastes, solvents, batteries, mercury	Organics	RCRA (Lindberg and Hartman 1999)
(a) Plan being revised to include 618-10 and 618-11 burial ground (groundwater beneath 300-FF-2 Operable			

Unit).

The 316-4 crib began receiving uranium-bearing waste solutions in 1948 and continued to receive nitrate, hexone, and organic wastes periodically through at least 1962. This site was investigated as part of a CERCLA limited field investigation for the 300-FF-2 Operable Unit (DOE 1996b).

In January 1999, a groundwater sample from well 699-13-3A was analyzed for tritium for the first time. The well is adjacent to the 618-11 burial ground. The tritium concentration was 1.8 million pCi/L, much higher than surrounding wells. Subsequent samples confirmed the high result and indicated that the burial ground was the source of tritium. An additional groundwater investigation is currently underway. Monitoring requirements will be specified in the revised Operational and Maintenance Plan for the 300-FF-5 Operable Unit, which will include contaminated groundwater beneath the 300-FF-2 Operable Unit (618-10 burial ground).

9.1.2 Groundwater Monitoring Requirements and History

Monitoring of groundwater levels and contaminant concentrations in the 600 Area were initiated in the 1940s. Water-table maps of the unconfined aquifer have been prepared at various times since 1944. The primary monitoring objective is to obtain data needed to track major groundwater contaminant plumes across the site as required by the *Atomic Energy Act of 1954* and its implementing orders. Wells not influenced by Hanford groundwater contamination are monitored to determine background groundwater quality. Wells across the Columbia River from Hanford Site operations are monitored to check for Hanford-derived contaminants that could be migrating offsite.

Additional wells were installed around the Central Landfill in 1986-1987 for RCRA (Nonradioactive Dangerous Waste Landfill) and Solid Waste Landfill monitoring. In 1993, DOE issued a proposal for an expedited response action for the 618-11 burial ground (DOE 1993). One well was installed in 1995 to support this action.

⁽b) Plan being revised.

9.2 Conceptual Model

Groundwater levels indicate that flow directions in the 600 Area west of the 200 West Area are generally from west to east. This reflects natural recharge and irrigation input into the upper Cold Creek and Dry Creek Valleys (Hartman 2000). Significant contamination is not present in this area, though nitrate is present in certain wells. It is inferred that irrigation is the primary source of nitrate in this area.

Movement of tritium and nitrate plumes and measurement of water levels provide a basis for inferring the direction of groundwater flow in the 600 Area across the central and eastern portions of the Hanford Site. The tritium and nitrate plumes, which originate in the 200 East Area and pass beneath the Central Landfill, indicate that the principal direction of groundwater flow is toward the southeast and east.

The rate of groundwater flow beneath the landfill is estimated to be on the order of 1.2 to 1.8 meters per day, based on site-specific hydrologic testing and the observed hydraulic gradients. However, estimates of groundwater velocity based on tritium and nitrate concentrations and tracer test results indicate groundwater transport rates of 6 to >30 meters per day. The lack of a detectable head difference in two well pairs located at the Central Landfill indicates that the vertical gradient within the upper portion of the aquifer is negligible.

The Central Landfill appears to have had little impact on Hanford Site groundwater, owing to minimal disposal of liquids at this facility. Associated groundwater monitoring consists primarily of measurement of RCRA indicator parameters (pH, specific conductance, total organic carbon, and total organic halides), though minor (below maximum contaminant level) contamination with chlorinated hydrocarbons exists.

Localized sources of contamination are present at the 618-11 burial ground (tritium), the 618-10 burial ground and the 316-4 crib (uranium and hydrocarbons).

Monitoring of the chemistry and hydraulic head in the upper basalt-confined aquifer is also conducted at the Hanford Site. The primary objective of this activity is to determine how much contamination is moving downward from the unconfined aquifer, and if contamination is moving offsite through this aquifer. The hydraulic gradient appears to be directed downward over most of the central portion of the Hanford Site, though the gradient is directed upward in the eastern portion of the site.

Water-level elevations north and east of the Columbia River are much greater than on the Hanford Site. The water-table elevation to the east of the Columbia River is currently 50 to 150 meters higher than on the Hanford Site. Groundwater flow in the unconfined aquifer system north and east of the Columbia River follows the bedrock structure and is toward the Columbia River. The water-table configuration in these areas largely reflects recharge from irrigation.

9.3 Monitoring Program

Site-specific monitoring networks in the 600 Area include the Central Landfill, the 618-10 and 618-11 burial grounds, and the 316-4 crib. The location of monitoring wells at the Central Landfill are

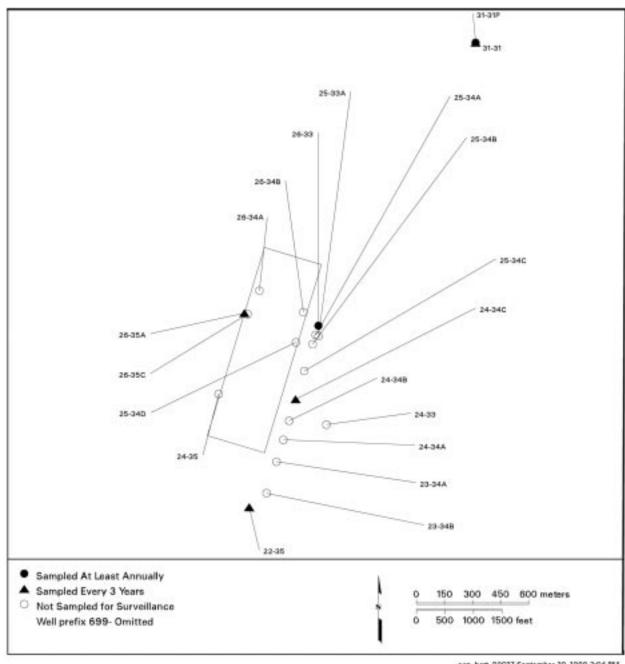
shown in Figure 9.1 (wells in the rest of the 600 Area were presented in Figure 4.8). Constituents monitored and sampling frequencies are presented in Appendix A.

Monitoring wells are maintained west of the 200 West Area and are sampled primarily for nitrate, which is probably related primarily to offsite agricultural activities.

There are six DOE wells located in the 600 Area north and east of the Columbia River, three of which have been used for contaminant-monitoring activities. Currently, monitoring of contaminant concentrations in this area is limited to well 699-42-E9B (shown in Figure 4.8).

Wells in the basalt-confined aquifer are sampled every three years. Well locations are shown in Figure 9.2.

The Operational and Maintenance Plan for 300-FF-5 Operable Unit is being revised to include contaminated groundwater beneath the 300-FF-2 Operable Unit, which includes the 618-10 and -11 burial grounds.



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Figure 9.1. Groundwater Project Well Locations: Central Landfill Vicinity

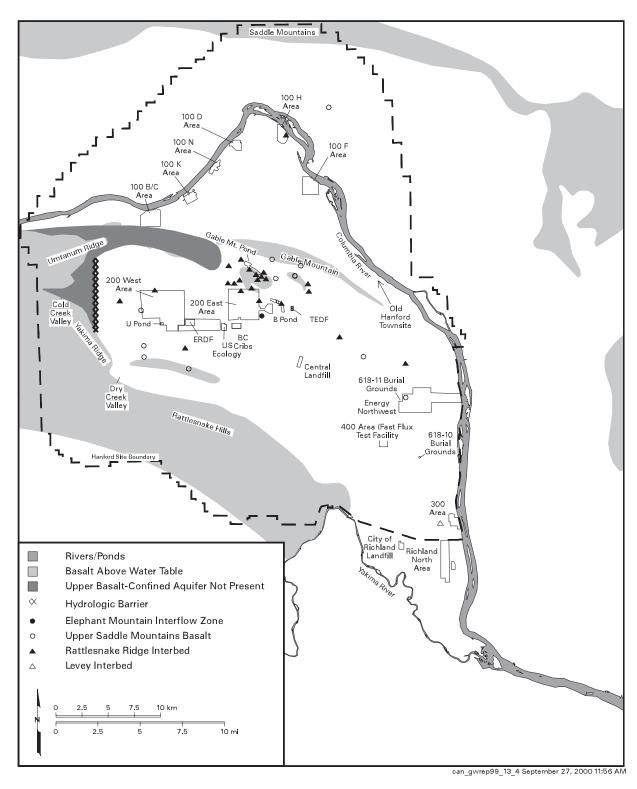


Figure 9.2. Wells Monitoring the Basalt-Confined Aquifer

10.0 Sampling and Analysis

10.1 Sampling and Analysis Protocol

Employees and subcontractors of Pacific Northwest National Laboratory sample wells for the ground-water project. Procedures for groundwater sampling, documentation, sample preservation, shipment, and chain-of-custody requirements are described in subcontractor manuals (procedures manual ES-SSPM-001¹) and in the quality assurance plan (ETD-012, Rev. 1² or latest revision). Samples generally are collected after three casing volumes of water have been purged from the well or after field parameters (pH, temperature, specific conductance, and turbidity) have stabilized. For routine groundwater samples, preservatives are added to the collection bottles before their use in the field. Samples to be analyzed for metals are usually filtered in the field so that results represent dissolved metals.

Procedures for field measurements are specified in the subcontractor's or manufacturer's manuals. Analytical methods are specified in contracts with laboratories, and most are standard methods from *Test Methods for Evaluating Solid Wastes: Physical/Chemical Methods* (EPA 1986). Alternative procedures meet the guidelines of SW-846, Chapter 10 (EPA 1986). Analytical methods are described in *Hanford Site Groundwater Monitoring: Setting, Sources and Methods* (Hartman 2000).

10.2 Quality Assurance and Quality Control

The quality assurance and quality control practices used by the groundwater project ensure the reliability and validity of field and laboratory measurements conducted to support these programs. The primary components used to assess data quality are accuracy, precision, and detection. Representativeness, completeness, and comparability may also be used. These parameters are evaluated through laboratory quality control checks (e.g., matrix spikes, laboratory blanks), replicate sampling and analysis, analysis of blind samples and blanks, and interlaboratory comparisons. Acceptance criteria have been established for each of these parameters. When a parameter is outside the criteria, corrective actions are taken to prevent a future occurrence. Quality control practices for the groundwater project and results for fiscal year 1999 are described in Hartman et al. (2000, Appendix B).

A scientist familiar with the hydrogeology of a particular location of a site or region reviews new data every two weeks. Staff conduct a more formal review quarterly according to a Pacific Northwest National Laboratory procedure to ensure the data are complete and representative. The review includes verification of the data in the Hanford Environmental Information System (HEIS) database, evaluation of data from field quality control samples (e.g., blanks, duplicates) and laboratory quality control samples. If the data review identifies suspect data, they are investigated to establish whether they reflect true

¹ Available from Waste Management Technical Services, Inc., Northwest Operations, Richland, Washington.

² Available from the Hanford Site Groundwater Monitoring Project, Pacific Northwest National Laboratory, Richland, Washington.

conditions or an error, according to Pacific Northwest National Laboratory's "request for data review" procedure. Groundwater data associated with out-of-range quality control data or identified as suspect during the technical review are flagged in the database.

11.0 Water-Level Monitoring

Water levels in the groundwater system are monitored on the Hanford Site primarily to help determine the direction and rate of groundwater flow. This information is used to interpret observed contaminant plume movements and to predict future plume movements. Other uses of water-level information include the identification of recharge and discharge areas, assessing the interaction between groundwater and surface water, assessing the interaction between aquifers or hydrogeologic units, calibration of groundwater-flow models, assessing the impact of liquid effluent disposal practices on groundwater flow, and optimizing monitoring networks.

McDonald et al. (1999) provides a list of wells used for water-level measurements, criteria for their selection, hydrogeologic units monitored, and describes procedures used to collect the data.

12.0 Data Evaluation

12.1 Data Management

Results of groundwater sampling and analysis are accessible in the HEIS database. Analytical results from all Hanford Site groundwater monitoring are stored in this common database, with the exception of some data collected for limited special projects that may not be directly comparable to standard data. The data are available to federal and state regulators for retrieval.

The HEIS programmers and HEIS data owners, including the groundwater project, ensure database integrity and data consistency through participation in the onsite HEIS technical advisory group and other ad hoc groups. The majority of data are loaded into the database from electronic files provided by the analytical laboratories under standard protocols. This minimizes data-entry errors and reduces the cost of data management.

As discussed in Section 10.3, a data validation and verification process results in flags and qualifiers based on quality control data and a technical review by a scientist. These flags are stored with the data in HEIS.

12.2 Compliance Issues and Data Evaluation

Data collected for the groundwater project are used to comply with a variety of requirements, including the *Atomic Energy Act of 1954* (and associated DOE Orders), RCRA, CERCLA, and WAC permits. After data are validated and verified, the acceptable data are used to interpret groundwater conditions at the site. Interpretive techniques include:

- Hydrographs: graph water levels versus time to determine decreases, increases, seasonal, or manmade fluctuations in groundwater levels.
- Water-table maps: use water-table elevations from multiple wells to construct contour maps to estimate flow directions. Groundwater flow is assumed to be perpendicular to lines of equal potential.
- Trend plots: graph concentrations of chemical or radiological constituents versus time to determine increases, decreases, and fluctuations. May be used in tandem with hydrographs and/or water-table maps to determine if concentrations relate to changes in water-level or in groundwater flow directions.
- Plume maps: map distributions of chemical or radiological constituents are ally in the aquifer to determine extent of contamination. Changes in plume distribution over time aid in determining movement of plumes and direction of flow.

- Contaminant ratios: can sometimes be used to distinguish between different sources of contamination.
- Concentration limits: contaminant concentrations are compared to drinking water standards, statistically-derived threshold values, or other concentration limits established in State or Federal regulations or agreements (Table 12.1).

12.3 Reporting

Results of Hanford Site groundwater activities are reported annually (e.g., Hartman et al. 2000). That report presents contaminant distribution maps, water-level maps, and concentration trend plots of contaminants and wells of interest; and meets the annual reporting requirements of RCRA and DOE Orders. CERCLA activities, including groundwater remediation and monitoring, are summarized.

Quarterly letter reports of RCRA data availability are submitted to the Washington State Department of Ecology. These reports include comparisons to critical mean values (background) for sites monitored for detection requirements, and summaries of recent results for sites monitored in assessment. Data from RCRA networks and the entire groundwater project are available on the HEIS database to the regulators.

Table 12.1. Compliance Issues and Methods of Evaluation

Requirement	Evaluation
DOE Order 5400.1	Compare groundwater concentrations to drinking water standards, derived concentration guides, and historical trends. Produce maps of contaminant distribution.
RCRA interim-status units	Indicator evaluation - Compare average downgradient concentrations of indicator parameters to background critical mean values.
	Assessment - Evaluate rate and extent of contamination (methods described in site-specific monitoring plans).
RCRA final-status units	Detection - Compare downgradient concentrations of contaminants of interest to baseline concentrations.
	Compliance - Compare downgradient concentrations to background, maximum concentration limits, or alternate concentration limits (methods described in site-specific monitoring plans and site permit).
	Corrective action - Track progress of cleanup and compare downgradient concentrations of constituents to background, maximum concentration limits, or alternate concentration limits (methods described in site-specific monitoring plans and site permit).
WAC-permitted units (216 permits)	Compare to conditions of permit.
CERCLA operable units (including performance assessment monitoring)	Compare concentrations to levels defined in records of decision, interim records of decision, or other agreements.

Certain conditions require reporting to DOE as unusual occurrences or off-normal events (DOE Order 232.1-1A). Those applicable to groundwater monitoring results include detection of contamination at the following levels:

- levels exceeding the derived concentration guide for a radionuclide in areas where this level has not previously been exceeded
- levels exceeding ten times the maximum contaminant level for hazardous constituents in areas where this level has not previously been exceeded
- levels above the drinking water standard or maximum contaminant level in areas where these levels have not previously been exceeded.

More stringent levels are set for areas adjacent to the Richland well field and/or south of Horn Rapids Road:

• 50% of the drinking water standard or maximum contaminant level for Hanford-derived contaminants, except for iodine-129 (maximum contaminant level is the detection limit) and tritium (10% of the drinking water standard in the vicinity of the Richland well field).

An occurrence report is also required if a RCRA groundwater assessment determines a facility has contaminated groundwater.

Reporting requirements for WAC-permitted facilities are described in their permits.

13.0 References

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